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1 **Tropical grassy biomes: linking ecology, human use and conservation**

2

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11 **Abstract**

12

13 Tropical grassy biomes are changing rapidly the world over through a coalescence of  
14 high rates of land use change, global change and altered disturbance regimes that  
15 maintain the ecosystem structure and function of these biomes. Our theme issue  
16 brings together the latest research examining the characterisation, complex ecology,  
17 drivers of change, and human use and ecosystem services of tropical grassy biomes.  
18 Recent advances in ecology and evolution have facilitated a new perspective on these  
19 biomes. However, there continue to be controversies over their classification and  
20 state dynamics that demonstrate critical data and knowledge gaps in our quantitative  
21 understanding of these geographically dispersed regions. We highlight an urgent need  
22 to improve ecological understanding in order to effectively predict the sensitivity and  
23 resilience of tropical grassy biomes under future scenarios of global change. With  
24 human reliance on tropical grassy biomes increasing and their propensity for change,  
25 ecological and evolutionary understanding of these biomes is central to the dual goals  
26 of sustaining their ecological integrity and the diverse services these landscapes  
27 provide to millions of people.

28 **1. Introduction**

29 Historically extensive across the global tropics, tropical grassy biomes (TGBs)  
30 are now changing rapidly through high rates of land clearance (1), increasing land use  
31 intensity (2, 3), woody encroachment (4) and disruption of the disturbance regimes  
32 (5, 6) that maintain ecosystem function. These biomes were the cradle of human  
33 evolution (7), and in our contemporary world, they support the livelihoods and  
34 wellbeing of over one billion people (8). With the population of Africa alone set to  
35 treble by 2050 (3), the continuing pace of climate change (9), increasing atmospheric  
36 CO<sub>2</sub> concentrations (9), and the increasing agricultural development of TGBs (3, and  
37 Estes et al., in this issue (10)), there is an urgent need to understand the unique  
38 ecology of these systems. TGBs, like tropical forests, are subject to a complex set of  
39 pressures as a result of human actions. However, unlike other biomes, the contrasting  
40 life forms and physiologies of the dominant C<sub>3</sub> woody plant species and grass species  
41 utilising the C<sub>4</sub> photosynthetic pathway sees the future of this biome linked, in a  
42 profound way, to the ever-rising atmospheric CO<sub>2</sub> concentration and the global  
43 political agenda to reduce these emissions. Further, TGBs have generally few policy  
44 and legislative mechanisms in place for their protection (11, see also the example  
45 provided by do Espirito Santo et al. (12)).

46 TGBs contribute 30% of global terrestrial net primary productivity and store  
47 15% of the world's carbon (13). While TGBs are less carbon dense than forests (by an  
48 order of magnitude or more), their productivity is such that large proportions of the  
49 carbon gained in a single year, are rapidly released back to the environment via fire,  
50 herbivory and human use (14, 15 and see analysis of this in Archibald and Hempson  
51 (16)). Indeed, the disequilibrium nature of TGB vegetation dynamics means that these  
52 biomes are highly sensitive to annual and decadal changes in environmental controls  
53 (14). The degree to which this dynamism will influence trajectories of vegetation  
54 change in grassy biomes into the future is unresolved. However, it is apparent that  
55 many intact savannas are now on a trajectory of increasing woody biomass, although  
56 the degree of gain varies regionally, with Australian savannas most stable over time  
57 (4).

58 Tropical grassy biomes first arose approximately 10 million years ago and  
59 expanded such that by two million years ago, tropical savannas and grasslands were a

60 dominant biome covering the tropics (17). Today, these biomes, cover in excess of  
61 20% of the global land surface. At the last glacial maximum, TGBs extended more  
62 widely throughout Asia, Africa and the Americas than today (18-20). The extent of  
63 these vast biomes has shifted with glacial - inter-glacial cycles in response to changing  
64 atmospheric CO<sub>2</sub> concentrations, rainfall, rainfall seasonality, temperature and fire  
65 (19, 21). Given that all of these aspects of our environment are now changing at  
66 unprecedented rates, extensive alterations in the distribution and dynamics of TGBs  
67 over the coming century will be inevitable and are likely already being observed (1, 4,  
68 22, see the analysis provided by Stevens et al. (23)).

69 The last decade has seen a revolution in our understanding of the evolution  
70 (17, 24, 25), antiquity (26-28), distribution (29, 30) and ecosystem dynamics of TGBs  
71 (14, 31), as well as their role in the global carbon cycle (32). Some of these advances  
72 have sparked controversies that are now active debates in the literature (e.g. ancient  
73 grasslands and afforestation policies (33-35); the existence or not of alternative  
74 vegetation states (29, 36, 37)). Indeed, insights and theory from savanna ecology have  
75 challenged long standing ecological assumptions of climate determinism in defining  
76 the limits of biomes (38).

77 Over the coming decade, we anticipate important in-roads will be made in  
78 reconciling the complex ecology and biogeochemical cycling of these geographically  
79 dispersed biomes via integration of remote sensing, modelling, ecology and evolution.  
80 However, it will be critical to incorporate the role of people in shaping and responding  
81 to changing ecosystem dynamics and function across this global region, as in the  
82 Anthropocene people will be increasingly important agents of landscape change,  
83 directly and indirectly influencing the environmental controls and ecological processes  
84 that structure TGBs from global to local scales.

85

## 86 **2. This Issue**

87 Tropical Grassy Biomes are expansive and changing rapidly, yet our capacity to  
88 predict trajectories of change in these biomes is limited, despite their importance to  
89 human livelihoods, biodiversity and biogeochemical cycling. In this issue, we highlight  
90 the need for integration among research related to the ecology and dynamics of these  
91 biomes: characterization and definition of tropical grassy biomes; complex ecology;

92 patterns and drivers of change; and, human use and ecosystem services. For the first  
93 time, analyses are presented on the biogeography and potential distributions of Asian  
94 savannas (39). Other significant steps forward in our understanding include: methods  
95 for characterizing ancient versus derived grassy biomes (40, 41), comparative data on  
96 the species diversity of TGB regions across the globe (42), an improved understanding  
97 of the complex ecology of herbivory and fire (16, 43) and the context dependent  
98 response of vegetation to global change (44), and finally, tools to examine tradeoffs  
99 in biodiversity, carbon and agriculture to aid land use planning and policy (10).

100

### 101 **3. Defining Tropical Grassy Biomes**

102 Tropical grassy biomes include C<sub>4</sub> grass dominated savannas and grasslands  
103 (following the definitions of (30, 45)). Definitions of tropical grassy biomes have  
104 historically been varied and fraught with problems. Functionally, TGBs are  
105 characterized by a grassy ground layer (generally dominated by grasses using the C<sub>4</sub>  
106 photosynthetic pathway - with a noted exception in Brazil (46) and Indochina (39)) and  
107 an overstorey varying from 0% up to 60 - 80% woody cover (45). The biota, depending  
108 on its biogeographic and environmental settings, is tolerant of any, or all of, fire,  
109 grazing and browsing (31). However, universally, the flora is shade intolerant, at least  
110 at the establishment phase, due to the open canopy overstorey (45). While the  
111 biodiversity value of these systems to-date has been typically overshadowed by that  
112 of tropical forests (26), Murphy et al. (2016) in this issue (42), illustrate the biodiversity  
113 value of TGBs, particularly of vertebrates and range-restricted species, and emphasize  
114 variation in diversity among the TGB regions (the South American region generally  
115 being the richest).

116 The disequilibrium nature of tropical grassy biome vegetation dynamics, with  
117 varying levels of woody cover, has consistently posed problems for the categorization  
118 of these ecosystems (8). This problem has been compounded by a focus on trees,  
119 rather than ground layer composition and function; for example, the Millennium  
120 Ecosystem Assessment focuses on drylands and forests (47, 48), but does not explicitly  
121 consider tropical savannas. While tropical rainforests have been mapped globally, no  
122 accurate global map of the tropical grassy biomes exists. The most widely used general  
123 vegetation map and classification scheme is Olson et al.'s (2001) ecoregions (49),

124 although this biome classification is problematic because it does not recognize some  
125 of the world's major savannas and grasslands, including those in Asia (e.g. India,  
126 Thailand, Burma) and Madagascar. Wide use of such maps for research, policy and  
127 conservation has the potential to have adverse impacts on landscape management  
128 and the perceived conservation value of these regions (e.g. conversion of TGBs,  
129 perceived as degraded land, for agriculture, see (50)). For the first time, Ratnam et al.  
130 (2016) focus on this issue in Asia (39) by reviewing the scattered literature on the  
131 distribution of Asian savannas and evidence for the antiquity and diversity of TGBs  
132 across this continent.

133 In many regions, including Madagascar, south east Asia and South America,  
134 grassy biomes have historically been considered either to be a degraded form of forest  
135 of anthropogenic origin created via tree clearing, burning and grazing, or a subclimax  
136 or secondary successional stage (28, 39). While true in some locations (see Veldman  
137 2016, this issue (40)), in the majority of areas this perspective is misplaced (34). A  
138 wealth of new information including dated phylogenetic analyses demonstrates the  
139 antiquity of both tree and grass species (and lineages) specialised to these biomes (25,  
140 27, 28). The presence of endemic plant lineages and species, as well as species with  
141 unique life histories and architectures, including forbs with large underground storage  
142 organs, are strong indicators of the antiquity of TGBs (35, 41).

143 The fauna of these regions also contains numerous endemic species  
144 specialized to open and grassy environments providing additional evidence for the  
145 origin and age of the tropical grassy systems. Fauna include species of granivorous  
146 birds (e.g. the Madagascan mannikin, *Lonchura nana*), a suite of grazing ungulates  
147 (e.g. the critically endangered Kouprey, *Bos sauveli*, from Cambodia, and the chital  
148 deer, *Axis axis*, from India) and a high diversity of small marsupials in Australia (42).  
149 Many of these species are endangered and threated with imminent extinction (51).  
150 Fire is a frequent disturbance in the TGBs and has been part of these systems for  
151 millions of years (17, 21); consequently the plants and animals they contain are  
152 generally adapted to its occurrence (25).

153 To date it has been difficult to distinguish ancient, old-growth grasslands and  
154 savannas from secondary systems given superficial similarities in structure. Here,  
155 Veldman (2016) and Zaloumis & Bond (2016), examining the Neotropics and South

156 Africa respectively, differentiate ancient and secondary systems, noting differences in  
157 species composition with the former particularly rich in forbs, many with well-  
158 developed underground storage organs that facilitate survival in seasonally dry  
159 climates with frequent fire. The challenge is to test the generality of these  
160 compositional characteristics across TGB regions.

161

#### 162 **4. Ecology**

163 The ecology of TGBs is complex by virtue of the numerous environmental controls,  
164 acting across different scales of influence, both directly and indirectly to structure  
165 these ecosystems (Figure 1). The last decade has seen a shift from a long-standing  
166 view of deterministic relationships among vegetation, climate and soils, focused on  
167 niche separation between trees and grasses for water use, to one that integrates niche  
168 separation (e.g., phenological, water use) with the controls of fire and mammal  
169 herbivory structuring vegetation via the restriction of woody plant growth (31).  
170 Archibald and Hempson (16) explore trade-offs in fire and mammalian herbivory  
171 across the African continent where realms of influence can change through space and  
172 time. Complementing this research is that of Anderson et al. (43) who examine spatial  
173 associations of African mammalian herbivores relative to body size and influences on  
174 ecosystem function relative to species composition. Both of these studies raise  
175 important questions about the function or dysfunction of TGBs in the context of  
176 changing disturbance regimes.

177 Integration of bottom up (e.g., climate and soils) and top down (e.g., fire and  
178 mammalian herbivory) controls in structuring TGB vegetation has significantly  
179 improved our process understanding of the dynamics and limits of these systems (14,  
180 52). But, it has also highlighted the degree to which contemporary dynamics of TGBs,  
181 from local to continental scales, are a function of historical contingencies (44, 53). In  
182 assessment of regional patterns and dynamics of vegetation, emergent patterns of  
183 woody cover can appear almost stochastic, due to the array of structural states  
184 possible for a given set of environmental conditions (29). At the heart of the current  
185 disagreement around alternate vegetation states prevalent in the tropical savanna  
186 and forest literature (all of state shifts between savanna and forest, grassland and  
187 savanna, and variation in tree cover within savannas) may be a lack of recognition of



188 both the role of contingency in influencing contemporary dynamics of TGBs, and that  
189 the relative role of environmental controls in structuring vegetation varies across  
190 savanna systems: i.e., some savannas likely exist due to soil barriers to woody plant  
191 growth, while others exist because of controls, such as prevalent fire, that also act to  
192 limit woody plant recruitment and growth. That is, the similarities in structure among  
193 TGBs (open canopied vegetation with a predominantly C<sub>4</sub> grassy ground layer) have  
194 led to an unfounded assumption in the literature that the processes regulating  
195 vegetation structure across these varied and geographically dispersed ecosystems are  
196 directly equivalent. Finally, the presence of numerous, well-documented, feedbacks  
197 structuring TGBs where the species composition can influence the strength and  
198 direction of effects (Figure 1; tree cover - fire; fire - grazing; grazing - browsing),  
199 combined with the importance of historical contingencies means that multiple states  
200 influencing both the limits and structure of TGBs are highly likely. In this issue, Oliveras  
201 and Malhi (54) examine the shades of green in our understanding of the processes  
202 structuring the limits of TGBs highlighting how biotic and abiotic processes operate at  
203 different scales and that nature of vegetation dynamics is context dependent.

204 Savanna vegetation dynamics have been shown to vary as a function of plant traits  
205 that aggregate from the individual to ecosystem level (14, 55). However, current  
206 model simulations generally represent TGBs as functionally identical, in contrast to  
207 ecological knowledge (although see Moncrieff et al. in this issue (44)). TGBs constitute  
208 a geographically dispersed set of regions, where the flora and fauna representing  
209 unique evolutionary and environmental histories (14). The relative importance of  
210 environmental controls in structuring these systems varies across these geographic  
211 regions, and relative to the environmental niche of each region (14). For example, the  
212 high rainfall Australian savannas dominated by tall, fast growing, narrow canopied  
213 evergreen *Eucalyptus* species are less sensitive to fire than the wide canopied  
214 deciduous *Brachystegia* and *Julbernardia* species that dominate a savanna region  
215 equivalent in area across southern Africa (56, 57). Thus, for a given set of  
216 environmental conditions, similar fire frequencies and intensities could produce  
217 different vegetation structures, and the difference in sensitivity to fire of these floras  
218 is highly likely underpinned by the functional traits of the plant species themselves  
219 (55). It is increasingly appreciated that the functional biogeography of TGBs has critical

220 implications for our capacity to determine the sensitivity and resilience of TGB regions  
221 to global change (e.g., Moncrieff et al., this issue (44)), and yet, our quantitative  
222 understanding of functional biogeography of TGBs remains limited. This information  
223 is needed as our capacity to predict future change will rely on a quantitative  
224 representation of the aggregation the traits that characterize these floras in  
225 influencing ecosystem dynamics and responding to environmental variation.

226 Across tropical grassy biomes, vegetation composition, woody cover and grass  
227 biomass are considered key determinants of ecosystem function. However,  
228 quantitative links between structure and function, ultimately, remain poor and there  
229 is no consensus of these relationships among TGB regions (8). Despite, the  
230 antagonistic dynamics between tree and grass dominance being central to savanna  
231 ecology, we retain a limited predictive capacity of vegetation structure. It could be  
232 argued that our current lack of knowledge about the physio-ecological responses of  
233 TGBs to global change is hindered by both the functional differences among the TGB  
234 regions and our weak quantitative understanding of the processes that structure  
235 vegetation due to the complexity of interactions and scales of feedbacks in operation  
236 (Figure 1). To aid the management of global change impacts for both people and  
237 biodiversity, we need to determine the relative sensitivities of savanna vegetation  
238 types to key environmental controls – CO<sub>2</sub>, water availability, and disturbance  
239 dynamics – and identify structural thresholds where critical ecosystem functions  
240 change.

241

## 242 **5. Drivers of Change**

243 Tropical vegetation is changing at broad spatial scales but there is a limited  
244 understanding of current trends. On one hand, rates of land use change are increasing  
245 (3, 11), and on the other, woody encroachment is widespread across savannas  
246 especially in Brazil and South Africa (in this issue, see Honda & Durigan, 2016 and  
247 Stevens et al. 2016). The extent to which drivers that enhance tree growth (e.g.,  
248 increasing atmospheric CO<sub>2</sub> concentrations [CO<sub>2</sub>]<sub>a</sub>), reduced disturbance, improved  
249 plant water use efficiencies), prevail over drivers of enhanced tree mortality (e.g.,  
250 reduced rainfall, increasing intensity of El Niño, increased temperature, increased  
251 harvesting) is unknown, but this is the key to the future management and integrity of

252 the biome.

253 Rising  $[\text{CO}_2]_a$  has long been hypothesised to be a key driver in the re-organisation of  
254 tropical vegetation, specifically in savannas where the contrasting life forms and  
255 physiologies of the dominant  $\text{C}_3$  trees and  $\text{C}_4$  grasses are expected to respond  
256 differently (58). While modelling of the proposed mechanisms underpinning shifts in  
257 the competitive interactions between  $\text{C}_3$  trees and  $\text{C}_4$  grasses is improving (i.e.  
258 increased plant water use efficiencies of  $\text{C}_3$  plant species, specifically woody plant  
259 species; carbon allocation and storage patterns that vary between life forms; reduced  
260 photorespiration in  $\text{C}_3$  grasses), demonstrating the potential for regional shifts in  
261 biome extent and woody biomass (59), there is a major gap in the experimental  
262 evidence of the responses of tropical plant species to altered  $\text{CO}_2$  concentrations,  
263 especially with regards to interactions with other environmental controls (8). In  
264 particular, dominant woody taxa in each savanna region have different life histories,  
265 allocation strategies, and architectures (14, 55, 60). Increasingly, functional traits are  
266 recognised as phylogenetically conserved (61), and differential responses to  $\text{CO}_2$   
267 would likely be expected relative to both ecological and environmental settings.  
268 Looking across a rainfall gradient and landuse types in South Africa, Stevens et al.  
269 (2016) report large increases in woody cover in just a few decades providing support  
270 for a global driver (23), while also noting the interaction with megaherbivores  
271 (elephants). Woody encroachment may provide carbon benefits, but will undoubtedly  
272 come at a biodiversity cost (62).

273 Tropical grassy biomes are characterised by seasonally dry and hot climates (30).  
274 While climates across this swath of the world are changing, particularly in terms of the  
275 frequency and intensity of El Niño drought events, disagreement among model  
276 predictions contributes to the lack of certainty for climate change predictions across  
277 tropical regions (63). Novel climates, in combination with rising  $\text{CO}_2$  will generate  
278 novel interactions among organisms, where small shifts in the season and timing of  
279 rainfall may have large consequences for the phenological cycles of plants and  
280 animals, and stark consequences for crop production (3). In contrast, small changes in  
281 total rainfall may be of limited consequence, where increasing  $[\text{CO}_2]_a$  will drive  
282 improvements in plant water use efficiencies (64). Temperature is assumed important  
283 in determining plant distributions and function primarily based on assumptions from

284 the Northern Hemisphere. Yet the importance of temperature in the dynamics of TGBs  
285 is poorly understood. The small body of research suggests if there is sufficient water,  
286 a warming climate may enhance plant success through improved germination (65, 66)  
287 and sapling growth rates (67), and an extended growing season (68).

288 Yet, across TGBs, rates of land use and cover change appear to exceed the effects  
289 of climate change (1, 11). With increasing global scarcity of lands for agriculture and  
290 increasing food demands (69), land use intensity is only likely to increase. There are a  
291 multitude of land use types across TGBs, many of which are context dependent, from  
292 shifting cultivation and grazing lands to commercial agriculture (see Ryan et al., in this  
293 issue). However, all affect the continuity of ecosystems and, some land use types more  
294 so than others (see Estes et al., in this issue). Increasing land use intensity and  
295 fragmentation disrupts disturbance regimes and vegetation dynamics (70), potentially  
296 amplifying encroachment by further reducing tree mortality. To date, 50% of the  
297 Brazilian cerrado has been transformed for agriculture, a rate of land use change  
298 roughly double that of the Amazon forest (71, 72). Land has historically been  
299 perceived as being of marginal agricultural value across TGBs (see Estes et al. and Ryan  
300 et al. in this issue). However, technical innovations in managing highly weathered  
301 tropical red soils and the breeding of suitable crop varieties have transformed  
302 agriculture in Brazil (3). This tropical agricultural revolution has been proposed as a  
303 viable development model for wetter African savannas (Estes et al. 2016 and (3)). do  
304 Espírito Santo et al. (this issue) document land abandonment and encroachment in  
305 secondary savannas of the cerrado (12), where this development policy has been to  
306 the detriment of the integrity of the system and where, globally, rates of  
307 encroachment are highest (4).

308 Finally, people also directly influence disturbance regimes at broad scales (6, 70).  
309 Active suppression of fire in the savanna regions of Asia and Brazil in particular has  
310 facilitated woody or weed encroachment (see Honda & Durigan, 2016 and (73)). The  
311 increasing extent of roads and fences act as fire breaks and also prohibit large scale  
312 animal movements (6). These changes to the major savanna processes of fire and  
313 herbivory, combined the effects of the poaching crisis of Africa and Asia, are likely to  
314 have profound consequences for ecosystem function (Archibald and Hempson;  
315 Anderson et al. in this issue).

316

317 **6. Human use and value**

318 Quantifying and understanding the value of TGBs to humans is challenging  
319 because in many regions, particularly in Asia, data on the value of TGBs to human  
320 livelihoods are limited. TGBs are sometimes described as “unused” or “degraded”,  
321 although these systems provide fundamental resources and ecosystem services  
322 supporting the livelihoods of the millions of people living in these regions. Further, the  
323 people of these regions are among the worlds poorest and most vulnerable (74), and  
324 global change will inevitably affect ecosystem services and resource availability.  
325 Perhaps more so than any other TGB region, direct use by local communities is  
326 greatest in Africa and Asia (see Ryan et al. 2016, this issue), given the urbanization of  
327 Latin America and the sparse population densities of Australia. Ryan et al. (2016, this  
328 issue) highlight the diversity and number of ecosystem services (supporting,  
329 regulating and cultural services) provided by TGBs (specifically in relation to southern  
330 Africa) that, to-date, have typically been either overlooked or considered at small  
331 spatial scales. Critical resources provided include food (wild fruits, tubers, nuts, edible  
332 insects, bushmeat), NTFPs for sale (e.g. honey, beeswax, insects), fuel (fire wood and  
333 charcoal), construction materials (e.g. thatching grass, timber), water, nutrient cycling  
334 and medicinal plants (Ryan et al. this issue).

335 Across Africa, Asia and even South America, fuel wood harvesting is a significant  
336 activity (see Woollen et al. and do Espirito-Santo et al. in this issue). Although the fuel  
337 wood crisis predicted in the 1970s has not materialized, projections surrounding levels  
338 of fuel wood sustainability are varied (75). In some regions (e.g. South Africa) fuel  
339 wood harvesting is considered sustainable due to regeneration after coppicing (76),  
340 possibly facilitated by CO<sub>2</sub> fertilization (e.g. (77)). Elsewhere, wood demand is  
341 anticipated to increase due to increasing populations and a switch to charcoal, which  
342 has an increasing international market (75). The implications of changing wood  
343 demands are raised by Woollen et al. (2016, in this issue (78)): wood for construction  
344 material is now traded off against wood for charcoal. How changes in wood resources  
345 in the context of global change will influence the integrity and functioning of TGBs  
346 needs urgent attention.

347 Water availability will change with growing populations, altered landscapes and

348 global change, and in the seasonally dry climates of these regions water availability is  
349 critical to both ecosystem dynamics and human use of landscapes, yet understanding  
350 of the nature of this change and the implications is restricted to a few regions (75).  
351 Honda and Durigan (2016, this issue (79)) provide, for the first time, estimates of  
352 rainfall partitioning in the cerrado and demonstrate how fire suppression, indirectly  
353 via woody encroachment, can reduce rain interception. Given many TGB regions are  
354 experiencing woody encroachment, this is a reminder of the functional role of  
355 disturbance in the provision of water resources. Across TGBs, afforestation (often of  
356 exotic species) is common (8), in part, to meet fuel wood needs; improved  
357 examination of the trade-offs between fuel wood and woody encroachment with  
358 water are desperately required. Further, much needed are examinations of the  
359 biophysical and biogeochemical consequences of woody encroachment and land  
360 cover change, such as albedo and nutrient cycling.

361 People and institutional structures can strongly influence and affect TGBs - their  
362 biodiversity, functioning and services (8). For example, in many areas declines in  
363 mammal species have been reported due to bushmeat hunting and poaching,  
364 savannas have become degraded through overgrazing, land abandonment, and  
365 afforested monocultures (80-82). The economic importance of the wildlife across  
366 TGBs, particularly in African and some Asian regions (and to a lesser extent Australia),  
367 is a unique economic advantage that contributes to the tourism industries of these  
368 regions. Major ecological changes driven by the global changes set out above,  
369 combined with human pressures could threaten these economic benefits. Human  
370 usage needs to be reconciled with ecological values in these areas, including  
371 biodiversity; human activity has already reduced the richest botanical savanna region,  
372 the Brazilian cerrado, to a series of, arguably, dysfunctional fragments. Finally, the  
373 climate change mitigation agenda could represent a threat to TGBs, as there is  
374 increasing talk of the need for “negative emissions” to meet the emissions targets set  
375 out at the Paris climate conference (COP21) in 2015 where inappropriate application  
376 of these targets could lead to afforestation of TGBs (34).

377

## 378 **7. Conclusions**

379 The pace and scale of change in TGBs is astonishing and will affect human livelihoods,

380 biodiversity, carbon and biogeochemical cycling. Yet, our capacity to predict the  
381 direction and extent of change, as well as the consequences, is currently limited.  
382 Research will improve our ecological understanding of TGBs, and it is clear that our  
383 capacity to effectively predict the sensitivity and resilience of each TGB region is  
384 dependent upon understanding the cascading and interacting effects of ecological,  
385 socioeconomic and global drivers across contrasting contexts. This is no easy task even  
386 from a purely ecological perspective as these systems are complex, ecosystem  
387 dynamics are context dependent and uncertainty surrounds the influence of global  
388 change drivers.

389         In order to better conserve and manage TGBs for the future, we must look  
390 beyond a simplified view considering only the tree layer, to a perspective that  
391 embraces the grassy ground layer and the unique functions associated with this. Only  
392 with this broader perspective will we be in a position to consider the range of  
393 trajectories and possible states that are likely across the different regions and how the  
394 influence of key drivers may vary. To apply the most appropriate conservation and  
395 management efforts in the right place, field studies are needed to characterize and  
396 determine the antiquity and value of TGBs more broadly. Additionally, field studies  
397 will help us understand how the multiple pathways for structural (and compositional)  
398 change links to the functioning of these biomes.

399         Experiments manipulating global change drivers (e.g. water availability,  
400 temperature, and CO<sub>2</sub>) will help unravel the complexities of savanna process and  
401 dynamics so we are in a stronger position to understand how different TGB regions  
402 may respond to future change. We need to work with land managers and politicians  
403 to ensure that processes critical to the healthy functioning of TGBs (i.e. fire, herbivory)  
404 are maintained. This will mean revisiting carbon mitigation initiatives, taking a more  
405 nuanced approach to applying REDD+ in TGBs, and ultimately recognizing TGBs are  
406 wholly different to forests in terms of ecosystem function. Many of the world's  
407 poorest live in the TGBs; it is therefore essential that the dual goals of sustaining the  
408 ecological integrity of this biome and supporting the people who live in these  
409 landscapes must not be viewed as competing demands.

410

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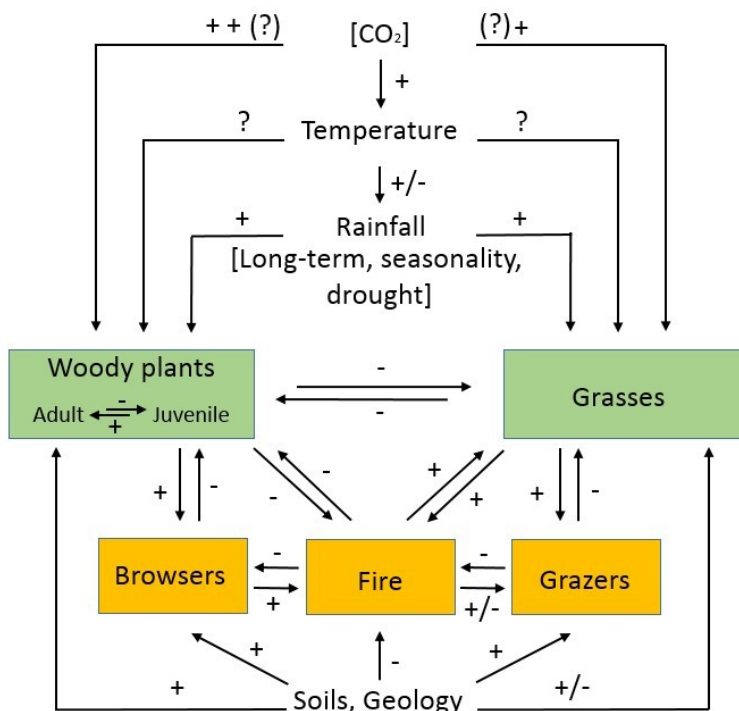
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644 Figure 1: *The complex ecology of tropical grassy biomes*. The network of interactions  
 645 governing the structure of tropical grass biomes, adapted and expanded from (83) and  
 646 of particular relevance is the extent of consumer centred feedbacks in structuring  
 647 these ecosystems. Direction of effects are indicated as positive (+) or negative (-)  
 648 based on literature (summarised in 14 and throughout this issue), and where the  
 649 literature is sparse or poorly reconciled, uncertainty of effects are indicated with (?).  
 650 With respect to interactions between CO<sub>2</sub> and plants, estimated effects are positive,  
 651 but there is uncertainty associated with the strength of interactions due to a lack of  
 652 experimental evidence and the potentially hysteretic effects of consumer centred  
 653 feedbacks. It must be noted that in some instances, species and their traits can modify  
 654 the direction and strength of effects, as is the case with interactions between fire and  
 655 grazing as outlined in (16) and relative to the ecological and environmental setting.  
 656 Hence, not all interactions are present across all TGBs and not all interactions are of  
 657 equal relative influence across ecological and environmental settings.  
 658



659