Storage of Carbon Dioxide in Saline Aquifers: Physicochemical Processes, Key Constraints, and Scale-Up Potential

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AU: This article has been accepted for publication with moderate revisions. The Reviewer notes:

This manuscript is about what its title says, and it is written by an impressive roster of specialists. The authors have done a good job of preparing an introductory paper for the readers of this journal, who though chemical engineers are unaware of the specifics of CO2 storage and of CCS in general. This is a most welcome and appreciated initiative! The review strives to strike the difficult balance between divulgating too much specialized knowledge to the "lay person with engineering background" that is obvious for an earth scientist, and hitting a good enough level of scientific depth.

The manuscript was reviewed by editor Michael Doherty who represents the chemical engineering reader who has no specific knowledge of the geochemistry/physics of sub-surface aspects of CO2 storage, and also by an expert in the field. Although we like the article very much, we have some specific suggestions to help the authors improve the manuscript so that it has the biggest impact on the intended audience.

In the bulleted comments below I (we) have reported some detailed suggestions.

Detailed comments:

- Introduction, first and third bullet points on page 2 about EOR, EGR and ECBM: contrary to storage in aquifers and in basalt these approaches lead to "enhanced" production of oil and/or gas. As a consequence, the overall impact on the climate is actually negative, because the emissions associated to the additional amount of oil and/or gas produced by injecting and storing CO2 exceed the emissions avoided by storing that amount of CO2. I recommend that the authors add a comment on this point. Is a climate-positive operation of EOR at all conceivable?

- Page 4, bullet point 1): I would say "pure" instead of "free-phase"; in fact, I do not understand the meaning of the expression "free-phase". Same in the title of the sub-section on the next page.

- Page 5, figure 1: this is an important figure and one that a chemical engineer can well understand, but only if it is provided in its entirety, i.e. also with the portion around the critical point, which is masked in this case by the gray box. Moreover, the two dashed lines are confusing and not really necessary as they indicate the sensitivity of the curve on the temperature uncertainty, i.e. a detail with respect to the general features of the curve, which are indeed very relevant. It is also important that the authors specify not only the temperature gradient that they assume but also the pressure gradient, which I guess it is hydrostatic but it is not said explicitly.

- Pages 5 and 6: it would be nice if, when variables are used in the text, they were typed in italic to be consistent with how they are typed in the corresponding equations.

- Page 6, figure 2: I think that this is really important, but not only in the context about the fluid dynamic regimes prevailing, but also to clarify the target geometry of the reservoir. This should really be explained to the reader of AR Chemical and Biomolecular Engineering. For instance, it is kind of surprising that the word "caprock" is used only twice in the manuscript, but it is never explained, nor it is explained that without caprock there is no CO2 storage in an aquifer. Clarifying the geometry and the features of an aquifer is also important to properly frame the necessary discussion on page 15 about "confined" and "unconfined" storage systems.

Commented [PR1]: Thank you!

Commented [PR2]: Agree this is an important clarification that is worth making - a sentence has been added.

Commented [PR3]: We have added a sentence to clarify the meaning of free-phase CO2 which is an establish concept (and which is not the same as pure CO2).

Commented [PR4]: Figure and caption have been modified to respond to these suggestions.

Commented [PR5]: Done

Commented [PR6]: Good point - A clarifying sentence on the sealing system (caprock) has been added.
Page 6, Figure 3: I do not really understand how I am supposed to interpret the figure; and the caption does not help much. I urge the authors to explain this better or to remove, if it is not essential.

Page 7, Figure 4: I recommend the authors to explain the figure better, and to have additional information in the caption; besides, the indications “r = 10 m” and “r = 1 m” in the figure are not clear.

Page 8, near the top: the sentence “the actual CO₂ dissolution rate is between 0.5% and 1% per annum” is not clear to me. Is it 1% of the amount injected in that year, or is it 1% of the cumulative amount injected?

Page 9, I do not know what Ma and Ka mean and could not find a definition when I searched the manuscript. Could you please define them? I suppose that readers will know what Gt means.

Page 12, in the first two paragraphs of the new section there are two lines, “Experience… equilibration” that are repeated twice. The first should be a mistake.

Page 13, Figure 6: what is the quantity “f” reported inside the figure?

Page 14, at the beginning of the section on capacity. In the four bullet points four capacities are defined: “theoretical”, “effective”, “practical”, and “matched”. Then at the end of the page authors talk about “investible resource”, which makes the whole paragraph a little confusing. Wouldn’t it be better to indicate to which capacity the expression “investible resource” refers to?

Page 15, first paragraph, there are two expressions that are not clear: “analytical analysis” and “structural closure.”

Page 15, about “injection and production strategies”. I think that in the Gorgon CCS project in Australia they produce formation water from the aquifer where CO₂ is injected, to my understanding in order to release pressure. Then water is reinjected in another formation where pressure management is less of a problem. I recommend that the authors mention this important and well-known case in this section on injection management.

Page 17, discussion around Eq. 8. It seems to me that the discussion here is too specialized for the journal’s readership. The authors discuss issues about interpreting seismic data without explaining the basics. I am sure that what the authors say is very relevant and important, but I am also sure that the readers will not get much out of it unless they explain the basics. For instance, seismic data from Sleipner are fantastic and not so well known, particularly in terms of the subtleties around their interpretation. I believe that the authors would do a better service to the community around AR Chemical and Biomolecular Engineering if they could show the Sleipner data, and explain and analyze them for the reader.

Page 19, about cost of monitoring. While the authors refer to Equinor data, I recall having heard from Simon O’Brien, who is in charge of storage for the Shell Quest CCS project, that they had planned a $5/ton monitoring program, and that they had reduced it to about $1/ton after a few years after realizing that some techniques were redundant and others were not necessary. I am not sure whether there is a usable reference about this piece of information, but it would be a nice confirmation of the figures provided by the authors.

Page 20, at the top. The authors maintain: “The major challenges for CCS scale up are not geological, but about financial incentives and business drivers.” I agree but, what about policy strategies and public perception issues? The authors might think that the latter are not important, but they might want to underline this. By the way, in the conclusions the authors talk
about “socio-economic challenges”, which is a bit broader than what they say here. It would be clearer if the two messages were better harmonized.

Page 20, figure 9: in the accompanying file this figure is indicated as figure 6. Moreover, I think that the authors should do a better job in explaining the curves and the symbols. Are the dots an extrapolation of the five Norwegian data points indicated in the text? Am I missing something? This seems to me to be an essential figure (together with the related discussion), and I believe that it should be crystal-clear.

Page 22, concluding bullet points. The first three are rather surprising to me because they seem to contradict many of the things that had been said in the article. How can the authors argue that gigatons of CO2 can be stored (see figure 9) when they say that “long-term stability and safety of CO2 storage” is a challenge (first bullet point)? How can they argue about low cost of monitoring as they do in the main text when they say that the “development of cost-effective monitoring solutions” is a key technology focus in the future (third bullet point)? I believe that I understand what the authors want to say, but I am not sure if they say it in the proper way. None of the authors claim affiliation with institutions numbered 4 and 5. If this is correct then these two institutions should be removed from the list.

Finally, we are surprised by the Disclosure Statement. The first and corresponding author, as well as three of the eight co-authors, work for Equinor. If Equinor has financial/commercial interests in CO2 storage then this should be disclosed for the sake of transparency. There is nothing wrong or incorrect about such interests—we deliberately invited Dr. Ringrose and his associates to write this article precisely because they have world-class expertise. However, readers should know how they got it.

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Storage of Carbon Dioxide in Saline Aquifers:
Physico-chemical Processes, Key Constraints, and Scale-Up Potential

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\textbf{Keywords}

\textsuperscript{**AU: Please add department**}

CO\textsubscript{2} storage, CO\textsubscript{2} capture and storage, CCS, geophysics, geochemistry, monitoring, dissolution

\textbf{Abstract}

CO\textsubscript{2} storage in saline aquifers offers a realistic means of achieving globally significant reductions in greenhouse gas emissions at the scale of billions of tonnes per year. We review insights into the processes involved using well-documented industrial-scale projects, supported by a range of laboratory analyses, field studies, and flow simulations. The main topics we address are: (a) the significant physico-chemical processes, (b) the factors limiting CO\textsubscript{2} storage capacity, and (c) the requirements for global scale-up.

Although CO\textsubscript{2} capture and storage (CCS) technology can be considered mature and proven, it requires significant and rapid scale-up to meet the objectives of the Paris Climate Agreement. The projected growth in the number of CO\textsubscript{2} injection wells required is significantly lower than the historic petroleum industry drill rates, indicating that decarbonization via CO\textsubscript{2} capture and
storage CCS is a highly credible and affordable ambition for modern human society. Several technology developments are needed to reduce deployment costs and to stimulate widespread adoption of this technology, and these should focus on:
- Demonstration of long-term retention and safety of CO₂ storage and;
- Development of smart ways of handling injection wells and pressure;
- Development of cost-effective monitoring solutions; and deployment of
- Development of CCS hubs with associated infrastructure. [**AU: Edit OK?**]

INTRODUCTION

Reduction in global greenhouse gas emissions is a key issue for modern human civilization. An essential part of any cost-effective solution to this challenge is long-term storage of CO₂ in deep geological rock formations, a process referred to as geological CO₂ storage (GCS), CO₂ capture and storage (CCS), or carbon dioxide sequestration. (The terms carbon sequestration and carbon storage are often used erroneously as shorthand for geological storage of CO₂ molecules.) In this review, we consider only the case of geological storage of CO₂ captured from man-made sources and stored in saline aquifers, as we consider this to be the dominant vehicle for realizing globally significant levels of CCS (1). Other potentially significant forms of GCS include:

- CO₂ storage as a part of CO₂-enhanced oil recovery (CO₂-EOR) projects (2, 3); or enhanced gas recovery (4, 5) projects (4, 5). This is where in these projects, CO₂ is injected into a partially depleted hydrocarbon field to recover a greater portion of the trapped oil or gas that remains in the reservoir rock pore space, by both increasing the reservoir pressure and reducing the viscosity of the oil. Such projects are typically the main route for CO₂ capture, utilization, and storage (CCUS) concepts, where utilization of CO₂ use acts as an economic incentive through owing to the revenue generated from production of additional hydrocarbons produced using the CO₂. [**AU: Edit OK?**][**AU: Abbreviations used fewer than two times in the main text have been removed throughout, per house style.**]

Since the produced hydrocarbons lead to further CO₂ emissions when combusted, CO₂-EOR projects have net positive CO₂ emissions to atmosphere, but can be viewed as a route towards future negative-emission CCS projects.
CO₂ storage in depleted gas and oil fields (6, 7).

CO₂ storage in coal formations (8) either via injection of CO₂ into unmineable coal seams or as part of enhanced coal-bed methane projects (analogous to CO₂ EOR[**AU: Spell out? Enhanced oil recovery?**]).

CO₂ storage in igneous rocks (especially basalts (9)), where enhanced rates of mineralization of injected CO₂ can occur.

The reason we focus on storage in saline aquifers is partly to limit the review but also because important insights into the processes involved in CCS have been gained via well-documented industrial-scale saline aquifer storage projects (10–12). We also argue that CO₂ storage in saline aquifers offers the main solution to achieving globally significant reductions in greenhouse gas emissions (13), while accepting that GCS in oil and gas fields and in basaltic rocks may play a significant role in some geographies.

It is also worth stressing that CCS is not considered as an alternative to other key solutions to achieving reductions in greenhouse gas emissions, including greatly expanded use of renewable sources of energy, societal and lifestyle changes, changes in land use, and more efficient use of energy overall. In most projections, CCS is anticipated to support between 10–15% of total cumulative emissions reductions through to 2050 (14). However, CCS is widely recognized as an essential part of the decarbonization process for modern human society, as it enables removal of CO₂ emissions from existing industrial and energy systems, as well as supporting negative-emissions solutions (15). Indeed, the International Panel on Climate Change reviews of global warming, climate change impacts, and mitigation activities (16, 17) have repeatedly shown that global warming cannot be realistically mitigated without CCS. It should also be stressed that engineered geological storage of CO₂ is a well-established technology with over more than 50 years of operational experience in CO₂ capture, utilization, and storage CCUS and 25 years of saline aquifer storage operations. Most notably, industrial-scale CCS using a saline aquifer started in 1996 with the Sleipner project in Norway (18).

The main questions we wish to address in this review are the following main questions:

1. What are the dominant physico-chemical processes that occur during saline aquifer storage?
2. What have we learned about the constraints on CO₂ storage capacity?
3. How can this experience be applied toward a strategy for global scale-up of CCS in order to meet climate mitigation targets?

We will address these questions using various field cases but will most frequently use the Sleipner case study as an illustration. This is arguably the best-documented and most-studied field case and certainly the longest-running saline aquifer storage project.

**PHYSICO-CHEMICAL PROCESSES**

After CO₂ is captured at the surface from a CO₂ capture plant, the CO₂ must be transported to a wellhead for geological storage. The CO₂ must then be compressed to be injected at a sufficient pressure to enter the geological formation at the in-situ pressure and temperature. This involves taking the CO₂ across the phase transition to be stored in the liquid or dense phase. Figure 1 shows an example that illustrates this phase transition for typical subsurface conditions in the North Sea. The fundamentally concept is that CO₂ is stored relatively deep (greater than ~800 m) to ensure that CO₂ remains in a dense form—either a liquid or as a supercritical phase. Regional differences in the geothermal gradient mean the critical depth for this phase transition varies. At intended storage depths, CO₂ has a density around 700 kg/m³ (slightly less dense than water), but at the same time has a viscosity which is more similar to that of hydrocarbon gases (CO₂ viscosity is around 0.06 cP at 1,500 m depth). This means that an appreciation of CO₂ storage involves a substance that is unlike the water or hydrocarbon resources that have traditionally been the focus of subsurface reservoir engineering. Put simply, CO₂ in the subsurface has a liquid-like density and a gas-like viscosity.

We have decades of experience of understanding, modelling, and monitoring CO₂ storage projects from both natural CO₂ stores and the growing collection of engineered storage sites. CO₂ Carbon dioxide has also been injected into reservoirs in many CO₂ EOR projects, and reservoir modelling of CO₂—brine—hydrocarbon systems is a relatively mature technology (19). Natural reservoirs of CO₂ derived from volcanogenic sources (20), notably several large accumulations in the Colorado Plateau and Rocky Mountains of the United States, can also be used to better appreciate the long-term processes acting on CO₂ retained in the subsurface over millions of years. Specifically, these natural analogues have been used to constrain rates of CO₂ dissolution rates of CO₂ in the brine phase (21, 22) and the rates of long-term CO₂ migration and leakage along faults (23).
We can group the physico-chemical processes that control the fate of CO₂ in a saline aquifer in terms of:

1. The fluid dynamics of free-phase CO₂ in a brine-saturated porous medium;
2. The dissolution of CO₂ into the aqueous (brine) phase; and
3. The formation of minerals by chemical reaction with the CO₂ introduced into the saline aquifer.

While these topics are more fully covered in several useful reviews and textbooks (24–27), and here we identify recent insights and provide an update on the current state of knowledge. Note that the term ‘free-phase’ refers to CO₂ that is not geochemically mixed with brine or minerals and is therefore potentially mobile as a separate fluid phase within the porous medium.

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**<COMP: PLEASE INSERT FIGURE 1 HERE>**

**Figure 1** CO₂ density versus depth diagram for typical subsurface conditions in the North Sea. The black line is the density function at the Sleipner location assuming a geothermal gradient of 35°C/km (+2°C/km; gray lines) and a hydrostatic pore pressure gradient. The CO₂ phase transition occurs at somewhere between 550-m and 750-m depth, depending on local temperature, supporting the generally assumed ‘critical depth’ of 800 m. Blue boxes show the relative volume occupied by CO₂ in the subsurface compared to surface volumes.

**Fluid Dynamics of Free-Phase CO₂ in a Brine-Saturated Porous Medium**

Injecting of CO₂ into a brine-filled permeable rock formation is part of a class of multiphase flow problems that have been extensively studied extensively (e.g., 24, 25). A two-phase CO₂–brine system is immiscible—the fluid are separated by a capillary interface. An important first approximation to the behavior of CO₂–brine systems is found by application of a set of dimensionless ratios that characterize the flow dynamics of two-phase immiscible flow systems (24, 28, 29). There are many ways of expressing these ratios, depending on the boundary conditions assumed; however, the most important ratios for CO₂ storage are the viscous/capillary ratio (\(N_{VC}\)) and the gravity/viscous ratio (\(N_{GV}\)), which for a 2D system [using the assumptions of...
(30) can be expressed as:

\[ N_{VC} = \frac{u_x \Delta x^2 \mu_{nw}}{k_{av} \Delta x (dP_c / dS_w)} \]

and

\[ N_{GV} = \frac{\Delta \rho \mu_{nw}}{u_x \Delta z \mu_{nw}} \]

Where \( u_x \) is the total flow velocity in the horizontal \((x)\) direction, \( \Delta x \) and \( \Delta z \) are the system dimensions, \( \mu_{nw} \) is the viscosity of the nonwetting phase (CO\(_2\)), \( k_{av} \) is the average permeability, \( \Delta \rho \) is fluid density difference, \( g \) is the acceleration due-owing to gravity, and \( (dP_c/dS_w) \) is the capillary pressure gradient as a function of wetting-phase saturation. [**AU: Variables in italic per equation, correct?**]

Since-Because the viscous force scales with flow velocity (a function of the applied pressure gradient), viscous forces will dominate close to the injection well (a few 400-500 meters[**AU: Correct?**]) but then decay outwards into the aquifer, where gravity and capillary forces will become increasingly important. The gravity force is controlled mainly by the fluid density difference but is also influenced by the vertical permeability and system anisotropy. Owing-Due to the high dependence of CO\(_2\) density on temperature, the in-situ density may be difficult to determine accurately for some settings. For example, the CO\(_2\) density at the Sleipner storage site varies from about approximately 700 kg/m\(^3\) near the injection well down to approximately around 350 kg/m\(^3\) at the top of the storage formation. Therefore, \( N_{GV} \) is variable, both spatially and over time. However, what is clear is that the interplay of viscous, gravity, and capillary forces results in an ‘inverted cone’ shape for the CO\(_2\) plume as it spreads into the aquifer and beneath a sealing caprock. A caprock is an informal term for a geological sealing system, typically comprising several mud-rock units which provide primary and secondary seals to the porous aquifer unit. This process is well understood in terms of guiding principles but is difficult to predict in detail in the real world (24–26). These concepts are usefully summarized in Figure 2 (based on Reference (31)), which also identifies the near-wellbore region where dry-out effects can occur (discussed in the next section).
Figure 2 Sketch of flow processes and flow regimes for CO₂ injection into an idealized storage unit (modified [**AU: with permission?**] from Reference (31)).

The rate and degree to which capillary and gravity forces become important away from the injection well are difficult to determine for two main reasons: (a) determining the changing value of the viscous/capillary ratio can be quite challenging, and (b) rock heterogeneity has a critical role which that is difficult to predict and model. For a homogeneous porous media, capillary forces only operate only at very small scales (at the pore-scale and up to around 0.2 m) and have little impact at larger scales. However, heterogeneous reservoir rock formations (especially the effects of lamination and bedding) mean that effects of capillary forces can be quite significant at larger scales (28, 32, 33). One important effect is referred to as heterogeneity trapping, whereby small-scale heterogeneities (e.g., low-permeability layering at the scale of 0.01–0.1 m) cause retention of the nonwetting phase owing to capillary forces (Figure 3). These effects have been documented in many laboratory studies (34, 35), and models of CO₂ storage systems that account for heterogeneity trapping demonstrate that a significant amount of CO₂ storage is likely to be in the form of residual CO₂ saturation (36, 37).

Figure 3 Trapping of CO₂ at high saturations upstream of a zone of high capillary entry pressure in an otherwise permeable sandstone rock. Heterogeneities in the form of low-permeability layers lead to more trapping than would be anticipated by the pore-scale capillary trapping mechanism alone and can contribute significantly to enhanced trapping of CO₂ in an aquifer (modified [**AU: with permission?**] from Reference (32)).

Another effect illustrated is that small-scale heterogeneities in the capillary pressure characteristics can significantly enhance or slow down the advancement of the plume (32, 38). CO₂ migration upward across pervasive sandstone bedding layers will be inhibited and sometimes trapped, as described above. When the CO₂ migrates laterally as a gravity current along-beneath a caprock, semi-parallel to layering, it will channel, sometimes in layers as small as centimeters in thickness, and the lateral migration rate will be enhanced significantly (as illustrated in Figure 4). Similarly, plume migration through isotropic heterogeneities of the type found in carbonate reservoirs, or through networks of less-pervasive bedding planes, will channel in a way analogous to a river finding a path of least resistance for
fluid flow, leading to enhanced plume migration.

Figure 4 Simulations of CO₂ injected into reservoirs that are heterogeneous and homogeneous, respectively, in their capillary pressure characteristics, but otherwise with equal the same average properties. The heterogeneous distribution of capillary pressure, $P_e$, is shown in the top image, where $r_x$ and $r_y$ refer to the horizontal and vertical correlation lengths of the model. Layered heterogeneity in the capillary pressure (middle image) results in CO₂ channeling and more rapid lateral migration of CO₂ (modified [**AU: with permission??** from Reference (33)).

Dissolution of CO₂ into the Aqueous Phase

The most important geochemical reaction for CO₂ storage in saline aquifers is dissolution of CO₂ into the brine phase. This process has an important role in stabilizing and securing long-term storage, but estimates of the effect vary enormously. We know that the process of molecular diffusion of CO₂ within a saline aqueous phase is a very slow process (25), and we also know that the convective mixing at the CO₂–brine interface is a much faster process, which is expected to dominate the rate of CO₂ dissolution. However, to initiate convective mixing, a diffusive boundary layer needs to develop and achieve a critical thickness before convection can occur. Using numerical analysis based on experimental data, Riaz et al. (Ref (39)) estimated that the critical time ($t_c$) for onset of convection and the characteristic wavelength ($\lambda_c$) of the convection cells are in the range of $10 \text{ days} < t_c < 2,000 \text{ years}$ and $0.3 \text{ m} < \lambda_c < 200 \text{ m}$. As with fluid flow and trapping, reservoir heterogeneity further complicates the dissolution problem. The presence of heterogeneity in the permeability field can either inhibit or enhance the dissolution rates, depending on the sedimentary architecture (40). On the other hand, in contrast, free-phase CO₂ channeling through small capillary heterogeneities dramatically increases the overall CO₂–brine interfacial area. [**AU: Edit OK??**] This, in turn, significantly enhances mass transfer into the aqueous phase, such that CO₂ dissolution rates can even approach the same order of magnitude as the injection rate itself (41). Thus, reduction in this large range in a priori estimates requires more detailed knowledge of the geological architecture and permeability.

For the Sleipner case, where for which we have a relatively good knowledge of the aquifer properties, as well as good monitoring data to constrain the growth and geometry of the plume, we can estimate that the actual CO₂ dissolution rate is between 0.5% and 1% per annum, or between 10% and 15% of the total cumulative injected mass after 20 years (42, 43).
Figure 5 shows example results from reservoir simulations of the Sleipner storage unit, with estimated ranges in the dissolved fraction. Here, the high-resolution 2019 Sleipner reference model grid (2 million cells, with vertical cell thickness of 2 m) was used with the E300 reservoir simulator package, where CO$_2$–brine mutual solubilities are calculated assuming fugacity equilibration between brine and CO$_2$ phase using the method and data from Reference [44]. The simulation results show that by the time of the 2013 time-lapse seismic survey, the dissolved fraction was between 10.6% and 12.6% of the total CO$_2$ mass injected (Figure 5a). This estimate is consistent with laboratory data ([45]) and within the upper bound of the dissolved mass fraction that can be estimated by inversion of gravity field survey data ([46]). For longer-term forecasting (Figure 5b), the predicted dissolved fraction is very dependent on the vertical permeability assumption. For a low vertical-to-horizontal permeability ratio ($k_v/k_h$) (red curve in Figure 5b), CO$_2$ tends to migrate much faster laterally during injection, thereby increasing the dissolution of CO$_2$ by increasing the contact area of the CO$_2$–brine interface. For a higher $k_v/k_h$ ratio (green curve in Figure 5b), initial dissolution is lower because the plume remains more compact and has a lower CO$_2$–brine interface contact area. However, after injection is stopped, the more compact plume (high $k_v/k_h$) continues to spread and promote further long-term dissolution. Convection-driven dissolution is not included in these simulations. Forecasting long-term dissolution rates therefore remains a significant challenge, although shorter-term rates can be constrained from site data and are expected to be approximately 10% of the total mass for the Sleipner case.

Figure 5 Example reservoir simulation models of the Sleipner storage site with estimated ranges in forecasts for the dissolved fraction. (a) Simulated dissolved fraction for the historical period—The green curve is the reference case with 0.6-Mtpa injection, and the red curve is for a 1-Mtpa rate; figures to the right give estimates at 2012: the maximum possible dissolved fraction is estimated from gravity survey data ([42]). (b) Long-term forecasts: The green curve is the reference case, with $k_v/k_h = 0.1$, and the red curve is a corresponding low-vertical-permeability case with $k_v/k_h = 0.0001$. For both cases, 9.06 Mt CO$_2$ was injected up to 2012, when injection was stopped and the plume was allowed to stabilize with continuing dissolution. Simulations were done using the E300 simulation package (CO2Store option, with assumption based on Reference [44]) and using the grid from the Sleipner 2019 benchmark model (co2datashare.org). Actual injection at the site is variable and lies between 0.6 Mtpa and 1 Mtpa and continues to the present day.
We can, however, use natural analogues to estimate the longer-term rates. Combined noble gas and stable carbon isotope analysis of gas samples from nine actively producing natural CO$_2$ reservoirs in the United States, Europe, and China indicate that dissolution of CO$_2$ is the dominant storage mechanism over geological time in both siliciclastic and carbonate reservoirs (47–49). These studies showed that up to 90% of initially emplaced CO$_2$ in contact with sampled wells had been lost to dissolution. Furthermore, this work highlights that mineral precipitation is a minor sink, even after millions of years of CO$_2$ storage. More recent work using these methods, identified that some 7 kt of the ~1.5 Mt (Million tonnes) of CO$_2$ injected into the Cranfield EOR field, (Mississippi, USA) (~approximately 0.2%) had dissolved into the groundwater over an 18-month injection period of injection (50). A significant additional proportion of CO$_2$ had also dissolved into the oil phase within the reservoir, enhancing recovery from the field.

This gas geochemistry approach has been further developed by two recent studies (51, 52) have further demonstrated this approach, which have estimated both the total mass of CO$_2$ dissolution and the dissolution rate within the Bravo Dome CO$_2$ reservoir (~New Mexico, USA). Sathaye et al. (53) used thermochronology to estimate the timing of CO$_2$ charge into Bravo Dome to be 1.2 to 1.5 Ma. Using a specially constructed reservoir model, they determined that the mass of CO$_2$ currently retained within the reservoir is ~1.3 Gt (Gigatonnes). Integrating this reservoir model with the previous geochemical measurements, allowed estimation of the total amount of CO$_2$ emplaced as ~1.6 Gt, where an estimated 366 ± 120 Mt (22 ± 7%) of this had dissolved. It is estimated that >40% of the CO$_2$ dissolution occurred during emplacement, with the remainder subsequently dissolving into the underlying aquifer. In one sector of the reservoir, the rate of CO$_2$ dissolution determined was 0.1 g/(m$^2$y), which exceeds the amount expected from CO$_2$ diffusion alone, implying that convective mixing of CO$_2$ and water had occurred.

In contrast, Zwahlen et al. (52) took an alternative approach who modelling noble gas and stable isotope diffusion profiles from the gas–water–contact through the gas column to obtain a much younger estimate of CO$_2$ emplacement within Bravo Dome of 14,000–17,000 ka years ago (54). This work also calculated the amount of CO$_2$ lost to dissolution within the field, producing a larger estimate of 506 ± 166 Mt, indicating a significantly higher dissolution rate of 48 ± 17 g/(m$^2$y) to 58 ± 20 g/(m$^2$y). While Although to date CO$_2$ dissolution rates have only been constrained from only a single
natural analogues to date, and vary considerably, the work highlights the potential of the geochemical methods involved in assessing the effectiveness of different CO$_2$ trapping mechanisms, particularly if tracers inherent within the captured CO$_2$ are used (54, 55). The work also emphasizes that a thorough understanding of the hydrogeological setting of prospective CO$_2$ stores is essential for accurate prediction of the long-term fate of the injected CO$_2$.

Mineralization of CO$_2$

Introducing CO$_2$ into a saline aquifer unit will modify the natural chemical balance and potentially cause dissolution or precipitation reactions. Carbon dioxide is a common component of the subsurface rock system (it is the most abundant subsurface fluid in the crust apart from water), occurring both as a dissolved component of aqueous fluids (groundwater) and as a free/mobile gas phase. The main sources of naturally occurring CO$_2$ are (a) from (i) volcanic systems, with the CO$_2$ being sourced from the deep mantle (56), and from (ii) gas generated from buried biogenic sources. The major natural accumulations of CO$_2$ in North America (e.g., Bravo Dome, New Mexico, and Sheep Mountain, Colorado), which are used as sources for CO$_2$ EOR projects, contain CO$_2$ of a predominantly originating from the mantle origin. CO$_2$ is also produced from a wide range of biologically-sourced systems, including decomposition of organic matter, methanogenesis (a by-product of methane-producing microbes), oil-field biodegradation, hydrocarbon oxidation, and decarbonation of marine carbonates.

When introducing CO$_2$ into a saline aquifer, the main question is how the additional CO$_2$ might modify or perturb existing chemical reaction processes. Will some of the CO$_2$ precipitate as minerals (usually carbonate minerals or clays minerals), or could some dissolution occur? Some general conclusions can be made based on geological data from natural analogues (57):

- When CO$_2$ is added to siliciclastic rocks, such as sandstones, as soon as the formation water is saturated with CO$_2$, the injected CO$_2$ will simply remain as a separate phase.
  Over centuries or longer, feldspar group minerals may react with CO$_2$ that has dissolved into the reservoir brine to form carbonates and clays (58).

- In the case of CO$_2$ injection into carbonates (or rocks with carbonate cements), some dissolution of carbonate minerals will occur, but again, as soon as the formation water becomes saturated with CO$_2$, the injected CO$_2$ will remain as a separate phase.
Experience from early CO$_2$ storage injection projects, such as Sleipner, In Salah, and Snøhvit, confirms that geochemical reactions are slow and relatively minor (59, 60), with virtually all the CO$_2$ remaining as a separate phase (liquid, gas, or dense) phase phase. In an analysis of data from a natural CO$_2$ reservoir (a CO$_2$-rich gas field), ref. Wilkinson et al. (61) showed that 70–95% of the CO$_2$ is present as a free phase, after tens of millions of years, with only approximately around 2.4% of the CO$_2$ stored in the mineral phase and a similar amount dissolved in the pore waters. The finding here is that although dissolution and precipitation reactions do occur when new CO$_2$ is introduced into the subsurface, the CO$_2$ quickly establishes a new chemical equilibrium with the in situ pore waters, following which reaction rates are very slow. CO$_2$ Carbon Dioxide dissolution into the brine phase can, however, be significant (see below).

When CO$_2$ is put into contact with clay minerals, the possible reactions and effects that can occur become rather complex. For example, in the case of CO$_2$ storage in shales (62), gas sorption can lead to significant CO$_2$ storage capacity in shale sequences. Geochemical reactions, such as dissolution of silicate minerals and precipitation of carbonate minerals, may also potentially have a measurable effect on the porosity, permeability, and the diffusion properties of shales.

For the case of sandstone saline aquifers (siliciclastic sedimentary systems), although some trapping of injected CO$_2$ as a mineral phase can occur, the reaction rates of reaction are very slow. Some dissolution of carbonate minerals may also occur, but again at very slow rates. An analysis of potential geochemical reactions at the Sleipner CO$_2$ injection site over a period of 10,000 years into the future (63) showed that geochemical reactivity of the Utsira sandstone is rather low, with mineral trapping making only minor contributions to CO$_2$ storage.

Another focus of geochemical-reaction analysis for storage has focused on the near-wellbore environment, where carbonate minerals may be formed when calcium hydroxide (Portlandite cement) reacts with CO$_2$. In a detailed study of geochemical modelling and experiments of brine-CO$_2$ reactions with wellbore cement, Carroll et al. (64) it was found that although important reactions can occur (precipitation of amorphous silica, calcite, and aragonite), the reaction of the hydrated cement with synthetic brine occurs rapidly (usually within 5–10 days). Geochemical modelling (65) to assess the potential impacts of the observed reactions indicated that these mineral products act to retard the rate of CO$_2$ migration, of CO$_2$ which might occur along
potential interfaces (e.g., cement–rock interface), implying that mineralization will tend to seal up potential leakage points in the near-wellbore environment.

In contrast, the case of CO$_2$ storage in basalts (and other basic igneous rocks) results in very high rates of mineralization, as demonstrated at the CarbFix injection site in Iceland, where approximately around 80% of CO$_2$ injected at a depth of 500–800 m within hot basaltic rocks was found to be carbonated as minerals within one year (66, 67). Although CO$_2$ storage in basalts is very different from saline aquifer storage, the insights may be relevant, especially where saline aquifer sandstone formations are interbedded with volcanic rocks, where enhanced mineralization of CO$_2$ can occur, as has been the case in Australian natural CO$_2$ reservoirs in the Otway Basin (68).

CONSTRAINTS TO REALIZING CO$_2$ STORAGE CAPACITY

Injectivity and Well Constraints

There are two fundamental constraints on CO$_2$ storage in a saline aquifer: the ability of the well(s) to inject CO$_2$ at the require rates and the ability of the aquifer formation to take the total CO$_2$ volumes. Geological limits on capacity are reviewed below, whereas the well constraints are reviewed here. The two are, however, closely interrelated. For a CO$_2$ injection well, there are two main pressure gradients to consider: (a) from the wellhead pressure, $P_{wh}$, to the bottom-hole pressure, $P_{bh}$, and (b) from the bottom-hole pressure, $P_{bh}$, into the saline reservoir formation, $P_{res}$. The first involves an increasing pressure gradient and the second a decreasing gradient, with $P_{bh}$ normally being the maximum pressure in the system. Thermal effects can lead to significant pressure variations, meaning that pressure estimation away from measurement points may be challenging but tractable using an equation of state (EOS) and reservoir simulation software.

Experience from operating wells shows that the flowing bottom hole pressure may take several hours to stabilize toward the shut-in bottom hole pressure owing to thermal equilibration. The Peng–Robinson and Soave–Redlich–Kwong equations are two commonly used EOS, which because they are relatively simple to implement (cubic equations), are widely used in modelling packages (69). The Span–Wagner EOS provides more accuracy for understanding detailed system behavior and complex mixtures (70) but is more demanding for numerical simulation.

Experience from operating wells shows that the flowing bottom hole pressure may take
several hours to stabilize toward the shut-in bottom hole pressure owing to thermal equilibration. Pressure gradients in the well-bore system can, to some extent, be controlled by appropriate choice of tubing diameter and use of wellhead or downhole chokes. For injection into formations with depleted reservoir pressure, heating of the CO$_2$ stream may be required to avoid transition into the vapor phase, as was undertaken at the K12-B test site in the Netherlands (71).

Assuming that the $P_{bh}$ can be controlled by the design of the well and surface compression facilities, the flow rate from the well into the formation can estimated using the radial Darcy flow equation, which, assuming a vertical well geometry, has the form (27):

$$q = \frac{2\pi k_{res} h_i (P_{res} - P_{bh})}{\mu \ln(r_i/r_e)}$$

where $q$ is the CO$_2$ flow rate, $k_{res}$ is the permeability of the rock formation, $h_i$ is the height of the injection well interval (the completion interval), $\mu$ is the fluid viscosity, $r_e$ is the effective radius of the reservoir unit, and $r_w$ is the radius of the well itself. The far-field formation pressure, $P_{res}$, is usually assumed to be constant, but could gradually increase for the case of injection into a confined aquifer (e.g., a small fault block) or could decrease over time in the case of hydrocarbon production from gas fields in hydraulic communication with the injection unit (72).

Figure 6 summarizes the likely pressure gradients in the vicinity of an injection well, showing with a possible well-bore damage effect shown. The Injectivity Index, $II$, (a ratio of flow rate to pressure gradient) may be strongly influenced by these well-bore damage effects, causing the $P_{bh}$ to be much higher than the expected pressure ($P^*_{bh}$) without well-bore damage effects.

Experience from several projects (e.g., Sleipner, Snøhvit, and Quest) reveals unexpected variability in the injectivity performance in the early phases of projects (26, 73). Reasons for reduced injectivity performance include: formation collapse near the well (Sleipner), formation of salt precipitates owing due to reaction of CO$_2$ with brines (Snøhvit and Quest), and the
migration of fine particles which plug the rock pores (possibly at Snøhvit). To illustrate the typical magnitude of these near-wellbore effects, in the case of Sleipner, the first months of injection witnessed a tenfold reduction in injectivity owing to near-wellbore formation collapse and corresponded to a rise in $P_{sa}$ of approximately around 20 bars (26). Whereas in the case of Snøhvit, the first months of injection showed a fluctuating reduction in injectivity caused by salt precipitation and pore clogging and corresponding with a rise in $P_{sa}$ of approximately around 50 bars (74). In both cases, well interventions were applied to resolve the problems, and subsequent injection returned to close to the expected levels. At Sleipner, a new completion interval with gravel and sand screens was applied (73, 75), and at Snøhvit, a methyl–ethylene–glycol (MEG) solution was added to the injection stream (74). Injectivity constraints are therefore potentially significant but are likely to be resolved as part of the early-phase well management and optimization process. However, in several cases, where the encountered formation permeability in a CO$_2$ appraisal well was significantly lower than expected and insufficient for injection to proceed (76). In such cases, hydraulic fracturing could be used to enhance injectivity (76), or the well may need to be abandoned in search of alternative injection horizons/locations.

**Trap Capacity and Pressure Limits**

The capacity of the intended geological storage units is one of the most critical and debated aspects of saline aquifer storage. Different types of capacity estimates can be summarized by the techno-economic resource–reserve pyramid, in which several stacked capacity terms can be differentiated (77):

- a theoretical capacity (the physical limit);
- an effective capacity (an estimate using cut-off criteria);
- a practical capacity (considering economic, technical, and regulatory factors); and
- a matched capacity (site-specific storage realized for specific CO$_2$ projects).

Typically, national storage resource mapping projects use a form of effective capacity (e.g., 78–80), whereas industrial and engineering associations are more focused on practical and matched capacity estimates as a basis for investment decisions (81). The capacity of the Utsira formation offshore Norway (an extensive shallow marine sandstone of Miocene age),
which hosts the Sleipner CO$_2$ storage project, has been much studied in terms of future storage potential. CO$_2$ storage capacity estimates for the Utsira Fm range between 1 and 60 Gt depending on assumptions made; (82), ranging from the exploitation of structural traps only (at the low end) to development concepts using multiple wells, residual trapping, and pressure management (at the high end). However, the investable resource (i.e., a matched capacity) in terms of currently known and accessible prospects within the Utsira is estimated at approximately around 0.17 Gt (62), which illustrates the challenge in going from a potential storage resource to an investable resource for project planning.

The underlying physical process which controls the efficiency CO$_2$ storage efficiency in saline aquifers is that injection of a low-viscosity, buoyant, nonwetting phase into a water-saturated porous medium is fundamentally an inefficient process. The ratio of the actual volume of CO$_2$ stored to the theoretical pore volume available is termed the storage efficiency, $\varepsilon$, and represents the cumulative effects of heterogeneity, fluid segregation, and sweep efficiency. Analytical analysis using multi-phase flow theory supported by empirical site data suggests that $\varepsilon$ is in the range of 0.005 to 0.06 (i.e., less than 6% of the pore volume), with values of 0.04 or 0.05 being typical assumptions for regional storage resource mapping projects (58–60). The estimate for the well-documented Sleipner case is that $\varepsilon$ had reached approximately around 0.052 (26) by the time of the time-lapse seismic survey of 2013. There are also several potential ways to increase $\varepsilon$ above 0.06, by using smart well placements to exploit the geology (84, 85) or by modifying the injection stream (86). Filling of a structural closure (i.e., a geological trap) could allow $\varepsilon$ to exceed 0.5 within the closure, although there is no documented demonstration of this to date.

Similar to the process of exploitation of hydrocarbon resources, in which various injection and production strategies are used to enhance the recovery, it is theoretically possible to increase the storage capacity of a given reservoir by applying some advanced injection techniques designed to control the movement of CO$_2$ in the saline aquifer. These strategies can collectively be called mobility control techniques and aim at stabilizing the CO$_2$ front in the reservoir. This can be achieved in various ways using techniques adopted from hydrocarbon production. Water-Alternating-Gas (WAG) injection is a well-documented technique used in oil production to reduce unstable fingering of the injected gas stream and to increase the sweep efficiency. A similar scheme can be used in CO$_2$ injection to achieve control over the movement...
of CO₂ plume movement by injecting slugs of modified CO₂ stream following cycles of pure CO₂ injection. The bulk properties of CO₂ can be modified in various ways by using chemical additives, such as polymers or nanoparticles, or by intentionally fluctuating the temperature (71).

There are several ways of enhancing the CO₂ storage capacity via brine production to relieve the pressure, an approach which has been implemented at the Gorgon CCS project in Australia. The disposal of produced brine has environmental and financial implications, but properly managed has the potential to enhance storage resources in future projects. For the case of CO₂ injection into a confined geological system (e.g., a fault-block with low-permeability barriers), the storage efficiency may be much lower, e.g., $g < 0.01$ (87).

However, most geological systems have imperfect seals, allowing some pressure dissipation, so that closed system models are overly pessimistic. Basic rock and fluid compressibility arguments can be used to show that storage sites need to be situated within fault blocks large enough to allow adequate pressure dissipation (e.g., a 5 Mt injection requires a gross rock volume of >2,500 km$^3$ for a sealed boundary case) (26). For real systems, 3D fault architecture at the basin scale is likely to lead to some points of pressure communication through zones with lower fault displacement or fault zones with sand-to-sand juxtapositions.

For such confined geological systems, the storage capacity depends very much on the pressure history. Depleted reservoirs (owing to previous hydrocarbon extraction) can allow for higher storage efficiencies owing to the lower average pressure when injection begins. The same is true for aquifers under continuous depletion. This can occur if the storage reservoir is in hydraulic communication with a producing hydrocarbon reservoir. While hydrocarbon extraction is taking place, the storage units can experience considerable depletion, depending on the rate of extraction and the degree of pressure communication. The Smølaheia saline aquifer system located in the east of the Troll field, offshore Norway, is an example of such a system (72). Figure 7 shows a cross-section through a geological model of the Smølaheia storage prospects along with dynamic flow simulation results showing the distribution of the CO₂ plumes after injection of 2.4 Gt CO₂ from four (hypothetical) injection wells located at the southern parts of the aquifer and completed in both the shallower pressure-depleted Viking group aquifer and the deeper Dunlin Group aquifers that are expected to remain mainly undepleted. A low-saturation plume is spread over a larger area in the depleted reservoir, whereas higher-
saturation and more localized plumes are simulated for the deeper formations at close to hydrostatic pressure (Figure 7b). Thus, while although pressure limits may constrain storage capacity for certain cases, hydraulic connection to surrounding aquifers is likely to allow pressure dissipation. Effects of previous and concurrent pressure depletion will require dynamic flow simulation but can significantly improve or enhance long-term storage capacity (72).

Figure 7 (a) Example cross-section through a geological model of the Smeaheia storage prospects (WNW-ESE section through the Gamma structural closure), with white arrows indicating the two main storage targets and orange arrows indicating the main pressure communication points described by Wu et al. (120) (Section is 22 km wide and 2.5 km thick, color tone shows porosity). (b) Example flow simulation model in which 2.4 Gt of CO₂ has been injected in both the pressure-depleted Viking group aquifer and the undepleted Dunlin Group aquifers. The CO₂ plume is shown after 100 years of injection. (Left image) Location of the four injectors at southern part of the aquifer. (Middle image) CO₂ plume in the depleted Viking group. (Right image) High saturation and localized plume in the deeper Dunlin Group. Simulations were done using the E300 reservoir simulator package.

Monitoring to Optimize and Confirm Successful Storage

Monitoring is important to establish a license to operate for CO₂ injection projects. The site operator must adhere to legal requirements to demonstrate that the CO₂ is safely contained in the subsurface. Legal frameworks typically include requirements that the monitoring should demonstrate that the CO₂ is migrating as predicted in the subsurface, that it is safely contained, and that there is no risk of negative impact on the environment. Establishing effective ways to monitor CO₂ storage projects has drawn a lot of attention over the last two decades, and there are now several best practice documents, reviews, and textbooks on this topic to guide future projects (e.g., (88–90)). Although many of the successfully applied methods were originally developed for petroleum reservoir monitoring, CO₂ storage monitoring additionally involves a unique set of challenges related to the physical properties of CO₂ in the subsurface and a wide set of concerns around ensuring safe long-term storage. There is widespread agreement that the most effective tool for monitoring subsurface CO₂ migration in the reservoir is the use of repeat seismic imaging (4D seismic), which has been successfully used at the Sleipner (91), Snøhvit (74), In Salah (92), Ketzin (93), Tomakomai (94), Quest (95), and Aquisstore (96) saline aquifer storage projects. Other important monitoring technologies include
time-lapse gravity surveys, time-lapse resistivity logging downhole, and use of natural and artificial geochemical tracers (97, 98).

A key objective in 4D seismic monitoring of CO$_2$-storage is the detection limit, to establish the minimum threshold thickness for a CO$_2$-layer. Both the thickness and the velocity of a CO$_2$-layer typically change during injection, and it is a challenge to discriminate between the two effects (or to estimate both simultaneously) from conventional stacked seismic data, particularly as long as the layers are below tuning thickness (99–101). Figure 8 illustrates this challenge. As CO$_2$ is introduced into the aquifer unit, the velocity decreases significantly (Figure 8a), setting up a strong amplitude contrast in the system. However, below the tuning thickness, it is not possible to discriminate between the top and base of the layer from seismic data; hence, the thickness is undetermined (Figure 8b). The nonmonotonic behavior of the velocity as a function of CO$_2$ saturation further adds to this complexity, as it introduces an uncertainty in the time-thickness transformation. Above the tuning thickness, it is possible to separate the top and base of the layer and improve velocity constraints (101). This underlines the need for precise and highly repeated time-lapse seismic data, but also other methods are required if the aim is to constrain the thickness and saturation change determination. Despite these challenges, the seismic monitoring of Sleipner has led to important insights into how the CO$_2$ migrates in the subsurface, and the site has been used for testing dynamic flow models (102, 103). In particular, the degree to which CO$_2$ migration in the storage domain is controlled by the gravity-viscous ratio (Equation 2) or is dominated by capillary forces (Equation 1) has been significantly improved (104, 105) using the Sleipner seismic imaging data sets for calibration. Although this question is not fully resolved, the flow system is clearly gravity-dominated, and understanding vertical migration paths and migration flow dynamics is the key remaining challenge (106).

<COMP: PLEASE INSERT FIGURE 8 HERE>

**Figure 8** Illustration of the challenges in detecting thin layers with unknown CO$_2$ saturation from seismic data. (a) A wedge model of a CO$_2$ layer with varying thickness and saturation, showing that amplitude changes depend both on the saturation dependent velocity and layer thickness, whereas velocity is strongly dependent on CO$_2$ saturation. The velocity model assumes a-homogenous fluid distribution in the pore space. (b) Example cross-section through the 2010 seismic amplitude data at Sleipner showing amplitude variation for the CO$_2$ plume in Layer 9 (top layer).
In the case of thin horizontal CO\(_2\)-layers, the use of long-offset data or repeated refraction-type seismic data is a useful complementary tool to complement to conventional 4D seismic data. When a seismic wave is propagating horizontally, the detectability increases because the wave spends more time in the thin CO\(_2\)-saturated layer. Typical examples of such waves are head-waves and diving waves. In full-waveform inversion (FWI), such waves play an important role in stabilizing the seismic inversion process (\(108, 109\)).

If the storage unit is less permeable and the injection pressure increases owing to low injectivity, there is a need to discriminate between fluid saturation and pore pressure changes (\(110\)). A case study from the Snøhvit field (\(111\)) shows that use of prestack time-lapse seismic data (or near and far offset stacks) is one way to resolve this issue. Another way to resolve this is to combine various geophysical methods, for instance, time-lapse seismic and time-lapse gravity (\(112\)). The development of accurate seabed gravimeters (\(113\)) is an important contribution to making this possible.

Fiber-optic-based monitoring systems are currently in rapid development and have already been successfully applied for storage monitoring, with demonstrations of downhole distributed acoustic (DAS)-sensing for time-lapse monitoring of CO\(_2\) plumes demonstrated at the Aquistore (\(72\)) and Quest (\(114\)) and projects. Use of downhole and surface downhole distributed acoustic DAS for 4D seismic monitoring has great potential for reducing the monitoring costs of monitoring, as has been recently demonstrated at the onshore injection projects Aquistore and Quest in Canada (\(72, 114\)).

Another important concept is the trigger survey philosophy, in which a basic routine monitoring strategy is established with additional survey options that are deployed only when an anomaly is detected requiring further verification is detected. This is a key strategy for in reducing monitoring costs. Furthermore, monitoring should ideally be considered as a beneficial activity ensuring an overall cost-benefit for the lifetime operation of the storage project. In a study of ways to optimize offshore monitoring, the typical costs of monitoring based on historical experience at Sleipner and Snøhvit were estimated to be of order €24/t (for a 2015 reference) (\(115\)). Although this cost could be potentially be reduced further, it is a small fraction of total project costs and will ideally pay for itself in terms of avoided costs of project stoppages or avoidable well operations.
STRATEGIES FOR GLOBAL SCALE-UP TO MEET CLIMATE MITIGATION TARGETS

Even though CCS is widely considered a proven technology—there are currently 19 large-scale CCS facilities are in operation, together along with a further 4 under construction, which together have an installed capture capacity of 36 Mtpa (116)—there is a significant scale-up of in CCS deployment which is needed to meet the stated ambitions for emissions reduction in the next three decades. [**AU: Edi OK?**] Carbon capture and storage [**AU: CCS**] is projected to provide 10–15% of total cumulative emissions reductions through 2050, requiring annual storage rates in 2050 in the range of 6–7,000 [**AU: 6,000–7,000**]Mtpa (13). And even though the recent International Panel on Climate Change IPCC report on global warming (16) presents a range of illustrative model pathways with differing levels of assumed CCS, all the pathways require a significant CCS component of CCS. Cumulative storage growth rates in CCS deployment of at least 9% (117) are required, and with peak injection rates of up to 40–60 Gtpa by 2100. The total geological storage resource base required is not expected to exceed 2,700 Gt of capacity in underground reservoirs and may be significantly less[**AU: smaller**] (117).

In developing a strategy to meet these CO₂ storage goals, it is useful to consider a continental-scale geological framework for future saline aquifer storage. An analysis of global offshore continental margins (13) demonstrates that ample storage resources are available, and that these resources are typically close to the major industrial hubs and mega-cities, which are typically commonly located near major rivers feeding suitable offshore sedimentary basins. The major challenges for CCS scale-up are not geological, but are about financial incentives and business drivers. Public perception factors also play an important role in both resisting or encouraging CCS as a climate mitigation measure. Some form of societal incentive for CCS is needed; with carbon taxes, tax rebates, emissions standards and infrastructure investment funds usually dominating the socio-political discourse (ref). The analysis of storage on offshore continental margins (13) suggests that approximately about 12,000 CO₂ injection wells will be needed globally by 2050 to achieve the Paris Agreement goals (2DS). By using historic petroleum well rates as a proxy for potential future regional CCS well deployment, characteristic build-up rates can be estimated. Figure 9 shows well build-up rates for an illustrative continental CCS cluster (based on the historic Norway well database). Approximately 5 such clusters would be needed to meet global CCS targets by 2050, with each cluster needing approximately around 2,000 wells.

200 wells by 2030 and 1,000 wells by 2040. In practice, it is more likely that approximately around 10–20 smaller CCS hubs will emerge, focused around major national industrial clusters. These projected CO₂ well rates are significantly lower than the historic petroleum industry drill rates, indicating that decarbonization via CCS is a highly credible and affordable ambition for modern human society. For reference, more than over 1 million hydrocarbon wells were active in the USA in 2014 (the peak year to date) (118). The costs of saline aquifer storage (not reviewed here) depend very much on the injection depth, geological setting, and dimensions of the project, with reported cost estimates in the range of €2–20 EUR/tonne (2009 prices) (119). Onshore projects are generally cheaper than offshore projects, and large-scale CCS hubs will likely be the most effective means of reducing costs toward the lower end of this range.

Figure 9 Characteristics of a continental CCS cluster with well build-out rates based on the historic Norway well database. Cumulative CO₂ estimate is based on empirical well data with mean (bold lines) and P10–P90 range (dotted lines) using methods explained in Reference (13). Abbreviations: 2DS, two-degree scenario; CCS, CO₂ capture and storage.

SUMMARY AND CONCLUSIONS

We have reviewed the current state of knowledge for CO₂ storage in saline aquifers, using available large-scale field observations supported by laboratory data. During the project lifetime (nominally approximately around 25 years), the CO₂ is primarily trapped as a free-phase within the brine-saturated porous medium, with fluid dynamics controlled by an interplay of viscous, gravity, and capillary forces controlling fluid dynamics. Plume dynamics are macroscopically controlled by the viscous-gravity ratio, but capillary forces at the pore-scale and bedding-scale result in highly episodic migration behavior. Dissolution of CO₂ into the brine phase occurs slowly but steadily (as a function of temperature and salinity) and is found to be in the range of 10–13% after 17 years of injection for the Sleipner case. Study of natural CO₂ reservoirs, which are analogous of long-term geological storage, shows that hundreds of Mt of CO₂ can be trapped by dissolution over geological timescales. However, current estimates of the dissolution rate cover a wide range, from 0.1 to 58 g/(m²y), and more complete constraint of this important parameter is an active area of research. The fractionation of CO₂ into mineral phases is extremely
slow in most saline aquifer settings, such that mineral trapping makes only a minor contribution to CO₂ storage, even over periods of thousands of years. This strongly contrasts with the case of CO₂ storage in basalts, where storage of CO₂ as a mineral phase can dominate.

The main constraints on CO₂ storage in saline aquifers are related to injectivity limits and the rock formation capacity. Injectivity challenges have been encountered in some projects but are generally solvable using established well-management and intervention technology. The total formation capacity for CO₂ storage is generally less than 6% of the available pore volume, owing to the inherent inefficiency of the fluid dynamics of a low-viscosity buoyant immiscible fluid entering a water-wet porous medium. Monitoring of injected CO₂ as it migrates as a plume away from the injection point using time-lapse seismic surveys has proven to be a highly effective method for guiding project operations and for demonstrating storage assurance (termed conformance and containment in permit regulations). Continuing advances in geophysical imaging, especially using low-cost fiber-optic sensing, means that CO₂ storage monitoring programmes are likely to become increasingly in accuracy at reduced cost. While concerns about possible CO₂ leakage are important to acknowledge and address, a wide set of geophysical and geochemical diagnostic tools are available to assess anomalies.

Pressure barriers and the size of the geological unit in hydraulic communication with the injection horizon can further reduce these capacity limits. Despite these physical limits to storage capacity, the numerous thick accumulations of porous sandstones in the world’s sedimentary basins (especially offshore continental margins), provide more-sufficient storage capacity for the required CCS deployment in the coming decades. CCS deployment needs to grow from the current level of 36 Mtpa to more than 6,000 Mtpa by 2050, with a ceiling on rates of 40–60 Gtpa before 2100 in order to meet the emissions reduction requirements implied by the 2-degree warming scenario. This growth in CCS activity requires a CO₂ injection well-drilling rate reaching approximately 12,000 wells by 2050—a drilling activity which is of magnitude smaller than historic petroleum drilling activities.

Development of CCS hubs focused around major industrial clusters and exploiting the storage resources available in the world’s sedimentary basins offers an efficient and low-cost route to globally significant reductions in greenhouse gas emissions.

Although the most important challenges for future scale-up of CO₂ storage are socio-
economic, several technology developments could prove vital in reducing deployment costs and for stimulating widespread adoption of this climate mitigation tool. Key technology focus areas for the coming decade include:

- **Demonstration** Further efforts to understanding long-term stability and safety of CO₂ storage, by better understanding of fluid migration behavior, the rate of progress towards plume stability, and the rate of dissolution in the brine phase;

- Development of smart and interactive ways of handling injectivity variations and formation pressure limits to enable optimal use of multiple storage units within sedimentary basins;

- **Further Development efforts on developing** cost-effective monitoring solutions for assuring storage site performance, identifying anomalies, and modifying injection operations if needed (including fiber-optic solutions, trigger-survey concepts, and smart analysis of continuous and repeat-survey data sets); and

- Development of CCS hubs with associated infrastructure (e.g., pipelines, wells, compressors, and control systems) to connect CO₂ capture points from major industrial clusters to multi-well storage systems in high-porosity sedimentary basins. Recommissioning of aging infrastructure may be a viable way to avoid decommissioning costs while lowering CCS costs.

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**DISCLOSURE STATEMENT**

The authors are not aware of any affiliations, memberships, funding, or financial holdings that might be perceived as affecting the objectivity of this review. The authors, representing a range of academic and industrial institutions, each contributed specific technical aspects to this review and are not aware of any affiliations, memberships, funding, or financial holdings that might be...
perceived as affecting the objectivity of this review. Four of the authors work for Equinor ASA, a commercial entity engaged in CCS, oil and gas projects and renewable energy projects. The views and opinions of authors expressed herein do not necessarily state or reflect those of our employers, host institutions or our national governments or any agency thereof. S.M.V.G. is supported by UKRI EPSRC grant no. EP/P026214/1, NERC grant no. NE/R018049/1, and Total E&P.

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