Global inventory of nitrogen oxide emissions constrained by space-based observations of NO₂ columns

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[1] We use tropospheric NO₂ columns from the Global Ozone Monitoring Experiment (GOME) satellite instrument to derive top-down constraints on emissions of nitrogen oxides (NOₓ ≡ NO + NO₂), and combine these with a priori information from a bottom-up emission inventory (with error weighting) to achieve an optimized a posteriori estimate of the global distribution of surface NOₓ emissions. Our GOME NO₂ retrieval improves on previous work by accounting for scattering and absorption of radiation by aerosols; the effect on the air mass factor (AMF) ranges from +10 to −40% depending on the region. Our AMF also includes local information on relative vertical profiles (shape factors) of NO₂ from a global 3-D chemical transport model (GEOS-CHEM); assumption of a globally uniform shape factor, as in most previous retrievals, would introduce regional biases of up to 40% over industrial regions and a factor of 2 over remote regions. We derive a top-down NOₓ emission inventory from the GOME data by using the local GEOS-CHEM relationship between NO₂ columns and NOₓ emissions. The resulting NOₓ emissions for industrial regions are aseasonal, despite large seasonal variation in NO₂ columns, providing confidence in the method. Top-down errors in monthly NOₓ emissions are comparable with bottom-up errors over source regions. Annual global a posteriori errors are half of a priori errors. Our global a posteriori estimate for annual land surface NOₓ emissions (37.7 Tg N yr⁻¹) agrees closely with the GEIA-based a priori (36.4) and with the EDGAR 3.0 bottom-up inventory (36.6), but there are significant regional differences. A posteriori NOₓ emissions are higher by 50–100% in the Po Valley, Tehran, and Riyadh urban areas, and by 25–35% in Japan and South Africa. Biomass burning emissions from India, central Africa, and Brazil are lower by up to 50%; soil NOₓ emissions are appreciably higher in the western United States, the Sahel, and southern Europe. INDEX TERMS: 0345 Atmospheric Composition and Structure: Pollution—urban and regional (0305); 0365 Atmospheric Composition and Structure: Troposphere—composition and chemistry; 0394 Atmospheric Composition and Structure: Instruments and techniques; 1640 Global Change: Remote sensing; KEYWORDS: emissions, GOME, nitrogen oxides, biomass burning, retrieval, remote sensing


1. Introduction

[2] Nitrogen oxide radicals (NOₓ ≡ NO + NO₂) originating from combustion, lightning, and soils largely control tropospheric ozone production [Kasibhatla et al., 1991; Penner et al., 1991; Murphy et al., 1993; Jacob et al., 1996]. Tropospheric ozone plays a key role in determining the oxidizing power of the atmosphere, is an important greenhouse gas, and is toxic to biota. “Bottom-up” inventories of NOₓ emissions,
based on limited knowledge of emission factors and extrapolation, are subject to substantial uncertainties [e.g., Streets et al., 2003]. We show here that “top-down” information derived from space-based observations of NO2 columns can reduce significantly the uncertainties in NOx emissions when combined with a bottom-up inventory.

[1] The Global Ozone Monitoring Experiment (GOME) instrument on board the European Remote Sensing-2 satellite provides the capability for continuous global monitoring of NO2 atmospheric columns through observation of solar backscatter [European Space Agency, 1995; Thomas et al., 1998; Burrows et al., 1999]. The satellite was launched in April 1995 into a Sun-synchronous orbit, crossing the equator at 1030 LT in the descending node. The GOME instrument observes the atmosphere in the nadir view with a surface spatial resolution of 40 km latitude by 320 km longitude in the forward scan, using a scanning mirror to measure three such scenes across the flight track. Global coverage is achieved every 3 days after 43 orbits.

[4] Several groups including our own have retrieved and interpreted tropospheric NO2 columns from GOME [Leue et al., 2001; Velders et al., 2001; Lauer et al., 2002; Martin et al., 2002b; Richter and Burrows, 2002]. The retrieval involves three steps: (1) determining total NO2 line-of-sight (slant) columns by spectral fitting of solar backscatter measurements, (2) removing the stratospheric columns by using data from remote regions where the tropospheric contribution to the column is small, and (3) applying an air mass factor (AMF) for the scattering atmosphere to convert tropospheric slant columns into vertical columns. The more recent algorithms [Lauer et al., 2002; Martin et al., 2002b; Richter and Burrows, 2002] explicitly remove an artifact thought to be introduced by the diffuser plate on the GOME instrument. Our previous AMF calculation described by Martin et al. [2002b] improved on earlier work by applying accurate surface reflectivities obtained from GOME measurements at appropriate wavelengths [Koelemeijer et al., 2003], accounting for scattering by clouds, and resolving the spatial and temporal variability in the vertical profile of NO2.

[5] Aerosols introduce an additional complication in the AMF calculation, enhancing or reducing the instrument sensitivity depending on the aerosol single-scattering albedo and its vertical distribution relative to the trace gas [Palmer et al., 2001]. Previous NO2 retrievals used a maritime aerosol distribution [Velders et al., 2001; Lauer et al., 2002; Richter and Burrows, 2002] or neglected the aerosols altogether [Leue et al., 2001; Martin et al., 2002b]. The present work improves on these retrievals by accounting for spatial and temporal variability of the aerosol, its multicomponent character, and its correlation with the NO2 vertical profile.

[6] We go on to exploit the GOME tropospheric NO2 column as a constraint toward improving NOx emission inventories. In a pioneering study, Leue et al. [2001] used the GOME observations, together with an assumed constant NOx lifetime of 27 hours estimated from GOME observations in continental outflow regions, to derive global NOx emissions by mass balance. Their NOx emission inventory showed major anomalies relative to bottom-up estimates, with larger emissions from Africa than from North America and Europe combined, likely due to errors in the AMF calculation and in the NOx lifetime. Clearly, one cannot ignore the information from bottom-up NOx inventories, which represent a considerable level of knowledge, and are broadly consistent with atmospheric observations when implemented in global models [i.e., Emmons et al., 1997; Horowitz et al., 1998; Thakur et al., 1999; Staudt et al., 2003]. Our inventory integrates information from bottom-up inventories and the top-down GOME observations, weighted by their respective errors, to optimize the use of the GOME data for constraining NOx emissions.

2. Global Ozone Monitoring Experiment (GOME) Tropospheric NO2 Retrieval With Aerosol Correction

[7] Our retrieval of tropospheric NO2 columns presented here extends the July 1996 retrieval of Martin et al. [2002b] to an entire year, and accounts for the effects of aerosols on the AMF. We focus on the period September 1996–August 1997 during which satellite data including Along Track Scanning Radiometer (ATSR) firecounts and the Total Ozone Mapping Spectrometer (TOMS) Absorbing Aerosol Index constrain biomass burning emissions of aerosols and trace gases, and the large fires in Oceania associated with the 1997–1998 El Niño have not yet begun [Duncan et al., 2003].

2.1. General Retrieval

[8] Following Martin et al. [2002b] we determine total slant columns of NO2 by directly fitting backscattered radiance spectra observed by GOME over the wavelength region 423–451 nm. The stratospheric column and instrument biases are removed by assuming that NO2 over the central Pacific is mainly stratospheric, subtracting the corresponding columns from the ensemble of GOME observations for the appropriate latitude, scan angle, and day of observation, and correcting the result for the small amount of tropospheric NO2 over the Pacific. The stratospheric column is taken to be zonally invariant; this assumption does not cause significant error in summer [Martin et al., 2002b], but is questionable north of 45°N in other seasons due to dynamical variability [Douglass et al., 2001].

[9] We apply the AMF formulation of Palmer et al. [2001] to convert the resulting tropospheric slant columns into vertical columns. This formulation computes the AMF as the integral over the tropospheric column of the relative vertical distribution of NO2 (shape factor) weighted by the local sensitivity to NO2 of the solar radiation backscattered to space (scattering weights). The shape factor for every GOME scene is specified from a global 3-D model simulation of that scene (Goddard Earth Observing System, (GEOS)-CHEM model, see Appendix A), while the scattering weights are calculated from the Linearized Discrete Ordinate Radiative Transfer (LIDORT) model [Spurr et al., 2001]. Martin et al. [2002b] improved the original AMF formulation of Palmer et al. [2001] to use local monthly mean surface reflectivity retrieved from GOME [Koelemeijer et al., 2003] and to account for cloud scattering in partly cloudy scenes using local cloud information also retrieved from GOME [Kurosu et al., 1999]. Here we further improve on the AMF formulation by accounting for extinction by aerosols as described in section 2.2 and by performing the radiative transfer calculation for each GOME scene rather than interpolating over a table of parameters. We exclude scenes in which more than 50% of the backscattered intensity is from the cloudy
sky fraction of the scene as determined from the radiative transfer calculation constrained with the local cloud information. This threshold removes scenes with greater than about 40% cloud or snow cover.

2.2. Aerosol Correction

We account for aerosols in the AMF calculation by representing their vertically resolved optical properties within the LIDORT model. Aerosol data retrieved from GOME are insufficient to constrain the AMF calculation due to lack of information about single scattering albedo [Martin, 2002]. Instead, we use monthly mean fields of aerosol mass concentration for September 1996–August 1997 from the Global Ozone Chemistry Aerosol Radiation and Transport (GOCART) model [Chin et al., 2000a, 2002; Ginoux et al., 2001]. These fields include sulfate, size-resolved mineral dust and sea salt, hydrophobic and hydrophilic black carbon, and hydrophobic and hydrophilic organic carbon. Extensive evaluations of the GOCART simulation with in situ and satellite observations have been presented in a number of papers [Chin et al., 2000b, 2002; Ginoux et al., 2001]. The same aerosol fields used here in the aerosol correction to the AMF are also used to account for the effect of aerosols on photolysis frequencies and heterogeneous chemistry in the GEOS-CHEM model [Martin et al., 2003]. Aerosol optical properties including scattering phase function, single scattering albedo, vertically resolved optical thickness, and local hygroscopic growth factors are specified for each aerosol type following Martin et al. [2003].

Figure 1 (left) shows the global distribution of the aerosol correction factors to the AMF for March and August 1997. Aerosols reduce the AMF by up to 40% over the biomass burning regions of India and central Africa, largely due to obscuration of the NO2 column by black carbon with smaller contributions from organic carbon and mineral dust. The TOMS aerosol index [Herman et al., 1997; Torres et al., 1998] and in situ measurements [Kuhlbusch et al., 1996; Chowdhury et al., 2001] show high concentrations of absorbing aerosols over these regions. Aerosols reduce the AMF by 10–20% in dusty regions such as the Sahara, the northern tropical Atlantic Ocean, and the Middle East. By contrast, aerosols increase the AMF by 5–10% over industrial regions such as the eastern United States where scattering sulfate and organic aerosols dominate.

Figure 1 (right) shows the AMF including the aerosol correction. Values are generally in the range of 0.8–1.5 over land and 1.5–2.5 over oceans, reflecting a larger relative contribution of the boundary layer to the tropospheric NO2 column over land. However, relatively high AMFs are found over continental regions that are arid (deep boundary layer and high surface reflectivity), intensely convective, or snow covered. The seasonal variation in the AMF over the high-latitude oceans arises largely from the seasonal variation in solar zenith angle. The seasonal variation in the AMF over land is less pronounced except for regions of biomass burning.

Martin et al. [2002b] showed that the total error in the retrieval of tropospheric NO2 columns over continental source regions is largely determined by the AMF calculation. They derived through propagation of errors an overall error of 53% on the AMF calculation including contributions from errors in surface reflectivity (28%), the NO2 profile (15%),
aerosols (30%), cloud cover (28%), and radiative transfer (10%). Our accounting for aerosols in the present work reduces the corresponding error by at least a factor of 2. The error from clouds is reduced to about 20% by retaining only scenes in which more than 50% of the backscattered radiation is from the clear-sky scene. The overall error in the AMF for each scene is now estimated to be 42%, and surface reflectivity is its single largest contributor. The error in the NO2 retrieval due to spectral fitting and determination of the tropospheric column is about \( 1 \times 10^{15} \) molecules cm\(^{-2} \) and dominates the overall retrieval error over the oceans and remote continental regions [Martin et al., 2002b].

### 2.3. Sensitivity of the Air Mass Factor to the NO2 Shape Factor

[14] Most GOME NO2 retrievals have assumed a globally uniform shape factor [Leue et al., 2001; Velders et al., 2001; Lauer et al., 2002; Richter and Burrows, 2002], which leads to bias in the retrieval [Martin et al., 2002b] and presents difficulty when comparing retrieved and modeled columns, as the model may have in general a different shape factor than assumed in the retrieval [Palmer et al., 2001]. (Comparing retrieved and modeled “slant” columns would only displace the problem.) Our approach has been to use local shape factors from the GEOS-CHEM global 3-D model of tropospheric chemistry, both to optimize the retrieval and to enable subsequent comparison of modeled and observed columns [Palmer et al., 2001; Martin et al., 2002b]. However, a concern is that the retrieval then contains model information that is not transparent.

Figure 2 illustrates the model contribution to the retrieval by comparing the AMF calculated for a globally uniform shape factor with that calculated using NO2 shape factors from the GEOS-CHEM model (the AMFs in this comparison do not include aerosol effects). Martin et al. [2002b] show an example of the contrast between model shape factors over land and ocean. The globally uniform shape factor assumes a uniform mixing ratio in the lowest 1.5 km, decreasing exponentially above with a scale height of 0.6 km, similar to that used in previous retrievals [Leue et al., 2001; Velders et al., 2001; Lauer et al., 2002; Richter and Burrows, 2002]. The AMFs calculated with the globally uniform shape factor are about a factor of 2 lower than those calculated using GEOS-CHEM shape factors over remote regions, where actual NO2 columns include a large contribution from the free troposphere. Previous reports of NO2 plumes over the Atlantic Ocean from biomass burning or lightning [Spichtinger et al., 2001; Richter and Burrows, 2002] could thus be biased high. Over industrial regions the AMFs calculated from the uniform profile are usually within 10% of the GEOS-CHEM values in summer, but up to 40% too high in March when the vertical mixing of surface emissions extends to less than 1.5 km altitude.

### 2.4. Global Distribution of NO2 Columns

[16] Figure 3 (left) shows the seasonally averaged tropospheric NO2 vertical columns retrieved from GOME. As mentioned above, scenes where clouds or snow dominate solar backscatter have been excluded from the average in order to reduce the retrieval error. The resulting bias is insignificant. Including these scenes would change the averages by less than \( 5 \times 10^{14} \) molecules cm\(^{-2} \), less than the fitting error. The NO2 columns over industrial source regions are lowest in the summer due to rapid loss by reaction with OH [Velders et al., 2001]. Seasonal enhancements from biomass burning are observed over northern Africa during December-January-February (DJF), India during March-April-May (MAM), and central Africa and South America from June to October. There are no evident enhancements along ocean ship tracks, but we have verified that this is not inconsistent with current ship emission inventories [Corbett et al., 1999], which imply relatively low columns when averaged over a GOME scene. Regions of negative NO2 columns over high-latitude oceans likely result from zonal variation in the stratospheric NO2 column. Noise from the South Atlantic anomaly [Heitzler, 2002] is observed near South America where GOME is exposed to increased radiation. Consistent with Martin et al. [2002b], we find no enhancements from lightning over tropical continental regions such as Central America and South Asia during summertime. Lightning makes little contribution (generally less than \( 4 \times 10^{14} \) molecules cm\(^{-2} \) with local maxima of \( 9 \times 10^{14} \) molecules cm\(^{-2} \)) to the NO2 column during daytime [Martin et al., 2002b; Edwards et al., 2003], and therefore we would not expect to observe it from GOME.

### 3. Construction of a NOx Emission Inventory

#### 3.1. Strategy

[17] We construct here an optimized NOx emission inventory by (1) inverting the GOME satellite observations of tropospheric NO2 columns with the GEOS-CHEM model...
and (2) combining the resulting top-down NO\textsubscript{x} emission estimates with a priori estimates from a bottom-up inventory, weighted by the relative errors in the two estimates. From maximum likelihood, we express the monthly a posteriori emission, \(E\) (atoms N cm\(^{-2}\) s\(^{-1}\)), and its relative (geometric) error, \(\varepsilon\), for a given location and month as a function of the a priori NO\textsubscript{x} emission, \(E_a\), the top-down NO\textsubscript{x} emission, \(E_t\), and the relative errors, \(\varepsilon_a\) and \(\varepsilon_t\), on each. We assume a lognormal distribution of errors on \(E_a\) and \(E_t\), so that maximum likelihood yields

\[
\ln E = \frac{(\ln E_t)(\ln \varepsilon_a)^2 + (\ln E_a)(\ln \varepsilon_t)^2}{(\ln \varepsilon_a)^2 + (\ln \varepsilon_t)^2}
\]  

(1)

**Figure 3.** Seasonal mean tropospheric NO\textsubscript{2} columns for September 1996–August 1997. (left) GOME retrievals. (right) GEOS-CHEM model results sampled at the GOME overpass time. We capped the colorbar to highlight spatial structure; local retrieved NO\textsubscript{2} columns can approach 1.5 \(\times\) \(10^{16}\) molecules cm\(^{-2}\) over industrial regions during winter. The GEOS-CHEM model is described in Appendix A; NO\textsubscript{x} emissions used in GEOS-CHEM are described in section 3.2.
Table 1. A Priori GEOS-CHEM NOx Emissions for September
1996–August 1997

<table>
<thead>
<tr>
<th>Source</th>
<th>Emission Rate, Tg N yr⁻¹</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fossil fuel combustion</td>
<td>23.7</td>
</tr>
<tr>
<td>Lightning</td>
<td>6.2</td>
</tr>
<tr>
<td>Biomass burning</td>
<td>5.6</td>
</tr>
<tr>
<td>Soils</td>
<td>5.1</td>
</tr>
<tr>
<td>Biofuels</td>
<td>2.2</td>
</tr>
<tr>
<td>Aircraft</td>
<td>0.5</td>
</tr>
<tr>
<td>Stratosphere</td>
<td>0.1⁺</td>
</tr>
</tbody>
</table>

⁹The cross-tropopause NOx flux is 0.5 Tg N yr⁻¹ (including 0.1 Tg N yr⁻¹ as NOx and 0.4 Tg N yr⁻¹ as HNO₃).

and

\[(\ln \varepsilon)^{-2} = (\ln \varepsilon_a)^{-2} + (\ln \varepsilon_x)^{-2}.\]  

We apply this method to land surface emissions only, excluding lightning. Emissions of NOx over ocean are very small and inferring them from the GOME data would be prone to excessive error. As discussed earlier, GOME NOx columns are insensitive to lightning NOx emissions (most of the lightning NOx in the upper troposphere is present as NO at 1030 LT of the GOME measurement [Ridley et al., 1996]). An inventory that includes lightning would have ambiguous attribution of surface versus upper tropospheric emissions.

3.2. A Priori NOx Emissions and NO2 Columns

[18] The standard NOx emission inventory used in the GEOS-CHEM model [Bey et al., 2001a; Martin et al., 2002a, 2003a] is used here as the a priori. “Land surface sources here include contributions from fossil fuels, biofuels, biomass burning, and soils; it excludes contributions from lightning, aircraft, or the stratosphere. Table 1 contains the annual global NOx emissions in GEOS-CHEM from all sources. Figure 4 (left) shows the spatial and seasonal variation of the land surface emissions. Anthropogenic NOx emissions are from the Global Emission Inventory Activity (GEIA) [Benkovitz et al., 1996] and are scaled to 1996–1997 as described by Bey et al. [2001a]. Soil NOx emissions are computed locally using a modified version of the algorithm of Ytanger and Levy [1995], as described by Wang et al. [1998]. Emissions from biofuels and biomass burning use the climatological combustion inventories of Lobert et al. [1999] for biomass burning and Yevich and Logan [2003] for biofuels; vegetation-specific emission factors are described by Staude et al. [2003]. Biomass burning emissions are scaled to 1996–1997 and are distributed monthly on the basis of satellite observations [Duncan et al., 2003].

[19] Evaluations of the GEOS-CHEM simulation with surface and aircraft observations for NO in different regions of the world have been presented in several papers [Bey et al., 2001a, 2001b; Fiore et al., 2002; Martin et al., 2002a, 2002b]. The observations are sparse but are generally reproduced by the model to within a factor of 2. There is no indication of model bias in the vertical shape factor. Observations for NO2 are even sparser but there is good confidence in the ability of models to simulate the observed NO/NO2 ratio [Bradshaw et al., 1999]. Figure 3 (right) shows the calculated tropospheric NO2 columns in the GEOS-CHEM model using the NOx emissions from Table 1 for the global distribution of monthly means. The model captures the seasonal enhancements over industrial regions during winter in the Northern Hemisphere and over South Africa. Most seasonal enhancements from biomass burning are larger in the model than observed by GOME. Mean model columns are 5% higher than retrieved by GOME and well correlated (r = 0.75, n = 82,960, p < 0.005).

[20] The top panel of Figure 5 and Table 2 show the specified error \(\varepsilon_a\) on the a priori inventory. We estimate the error in the fossil fuel and biofuel components of the a priori GEOS-CHEM inventory by comparison with the recently developed EDGAR 3.0 NOx emission inventory for the year 1995 [Olivier and Berdowski, 2001]. The local ratio of GEOS-CHEM to EDGAR fossil fuel and biofuel emissions is taken as the relative error in these emissions; values are about 1.5 over much of the eastern United States and Europe. Annual global errors in biomass burning and soil NOx emissions are about a factor of 3 [Intergovernmental Panel on Climate Change, 2001], and we adopt these values here. Local monthly errors may be much higher, but an error of a factor of 3 on the a priori is typically large relative to the error on the GOME top-down constraint so that accurate error specification is not important.

3.3. Top-Down Estimate of NOx Emissions From GOME NO2 Columns

[21] Top-down inference of an NOx emission inventory \(E_t\) from the GOME NO2 columns by mass balance [Leue et al., 2001] requires at minimum three pieces of information: the retrieved tropospheric NO2 column \(\Omega_t\), the ratio of tropospheric NOx to NO2 columns, and the NOx lifetime \(\tau_{NOx}\), against loss to stable reservoirs. We obtain the latter two pieces of information from the GEOS-CHEM model. Mass balance for the NOx column then yields a linear relationship between \(E_t\) and \(\Omega_t\):

\[E_t = \alpha \Omega_t,\]  

where the linear coefficient \(\alpha = (\Omega_{NO}/\Omega_{NOx})/\tau_{NOx}\) is determined by the GEOS-CHEM chemical mechanism and is to a first approximation independent of NOx. The definition of \(\tau_{NOx}\) is complicated by rapid exchange of NOx with peroxyacetyl nitrites (PANs). We derive \(\alpha\) here from the simulation with a priori emissions as \(\alpha = \frac{E_t}{\Omega_t}\) where \(\Omega_t\) is the NOx column from that simulation. This approach has the advantage of yielding an effective \(\tau_{NOx}\) that accounts for the complexities in NOx chemistry and a \(\Omega_{NO}/\Omega_{NOx}\) ratio that accounts for free tropospheric sources as represented by the GEOS-CHEM model.

[22] In practice, \(\tau_{NOx}\) is mainly determined by the oxidation of NOx to HNO3 in the boundary layer. Horizontal transport of NOx over this NOx lifetime, which smears the local relationship between NO2 columns and NOx emissions, is a source of error in the application of equation (3). The NOx loss to HNO3 in the model is controlled largely by the OH + NO2 reaction in the tropics and in the extratropics during summer, and by N2O5 hydrolysis on aerosol surfaces during extratropical winter. Figure 6 shows the zonal mean lifetime of NOx against oxidation to HNO3 in the continental boundary layer in different seasons. The lifetime is 3–10 hours, except in extratropical winter where it increases to 1–2 days.
Figure 4. Seasonally averaged NO\textsubscript{x} emissions from land with $2^\circ \times 2.5^\circ$ horizontal resolution: a priori (left), a posteriori (b, e, h, and k), and difference between the two (right). Only surface emissions are included; nonsurface emissions from lightning, aircraft, and the stratosphere are excluded. We capped the colorbar to highlight spatial structure; values as high as $8 \times 10^{11}$ molecules cm$^{-2}$ are found in some industrial regions.
corresponding smearing length scale [Palmer et al., 2003] is ~100 km, on the same order as the GOME pixel size and the GEOS-CHEM model resolution; therefore we neglect smearing in our analysis. A more sophisticated inversion that accounts for NO transport would improve the inference of NO emissions from GOME observations during winter.

Figure 5. Relative annual mean error in the NO emission inventories. The relative error factor is the geometric standard deviation of the error. The a posteriori error is calculated from the a priori and GOME errors using equation (2). The colorbar is capped to highlight spatial structure; local error factors from GOME are very large over regions where retrieved tropospheric NO columns are below the detection limit of $1 \times 10^{15}$ molecules cm$^{-2}$ (Figure 3).

Table 2. Regional NO$_x$ Emissions (and Relative Errors)$^a$

<table>
<thead>
<tr>
<th>Region</th>
<th>A priori</th>
<th>EDGAR$^b$</th>
<th>GOME (Top-Down)</th>
<th>A Posteriori</th>
</tr>
</thead>
<tbody>
<tr>
<td>United States</td>
<td>6.5(1.8)</td>
<td>6.1</td>
<td>8.3(1.7)</td>
<td>7.1(1.4)</td>
</tr>
<tr>
<td>Europe</td>
<td>5.6(2.0)</td>
<td>7.6</td>
<td>5.8(1.7)</td>
<td>5.9(1.5)</td>
</tr>
<tr>
<td>East Asia$^c$</td>
<td>4.5(2.1)</td>
<td>5.1</td>
<td>4.7(1.7)</td>
<td>4.7(1.5)</td>
</tr>
<tr>
<td>Northern Africa</td>
<td>2.7(2.5)</td>
<td>2.7</td>
<td>4.0(1.9)</td>
<td>3.5(1.6)</td>
</tr>
<tr>
<td>South America</td>
<td>3.5(2.4)</td>
<td>2.9</td>
<td>2.2(2.7)</td>
<td>3.1(1.9)</td>
</tr>
<tr>
<td>South Asia (MAM)$^d$</td>
<td>1.4(2.3)</td>
<td>0.73</td>
<td>0.67(2.0)</td>
<td>0.89(1.6)</td>
</tr>
<tr>
<td>South Asia (rest of the year)</td>
<td>2.3(2.0)</td>
<td>2.2</td>
<td>1.4(2.3)</td>
<td>2.1(1.7)</td>
</tr>
<tr>
<td>North Asia$^e$</td>
<td>2.4(2.1)</td>
<td>2.0</td>
<td>2.9(2.0)</td>
<td>2.6(1.7)</td>
</tr>
<tr>
<td>Central Africa</td>
<td>2.4(2.8)</td>
<td>1.5</td>
<td>2.3(1.9)</td>
<td>2.2(1.7)</td>
</tr>
<tr>
<td>Middle East$^f$</td>
<td>1.5(2.0)</td>
<td>1.3</td>
<td>2.0(1.8)</td>
<td>1.7(1.5)</td>
</tr>
<tr>
<td>Australia$^g$</td>
<td>1.4(2.6)</td>
<td>1.4</td>
<td>1.1(1.0)</td>
<td>1.4(2.0)</td>
</tr>
<tr>
<td>Central America</td>
<td>0.85(2.3)</td>
<td>0.60</td>
<td>0.60(2.4)</td>
<td>0.75(1.8)</td>
</tr>
<tr>
<td>Japan</td>
<td>0.47(2.9)</td>
<td>1.0</td>
<td>0.74(1.6)</td>
<td>0.64(1.5)</td>
</tr>
<tr>
<td>South Africa</td>
<td>0.45(2.1)</td>
<td>0.54</td>
<td>0.73(1.9)</td>
<td>0.57(1.6)</td>
</tr>
<tr>
<td>Canada</td>
<td>0.55(2.0)</td>
<td>0.83</td>
<td>0.48(2.9)</td>
<td>0.56(1.8)</td>
</tr>
<tr>
<td>Total</td>
<td>36.4(2.0)</td>
<td>36.6</td>
<td>38.0(2.0)</td>
<td>37.7(1.6)</td>
</tr>
</tbody>
</table>

$^a$Units are Tg N yr$^{-1}$. Relative errors represent the geometric standard deviation about the expected value. Region boundaries follow standard convention. Values are annual means except for South Asia. They include all land surface sources (fossil fuel combustion, biofuels, biomass burning, and soils) and exclude nonsurface sources (lightning, aircraft, and the stratosphere).

$^b$The EDGAR 3.0 inventory only includes industrial, biomass burning, and biofuel emissions. We add the GEOS-CHEM soil NO$_x$ emissions here for consistency.

$^c$East Asia here includes China and Korea, but not Japan.

$^d$Totals for South Asia are shown separately for March-May (biomass burning season) and the rest of the year. South Asia here includes India and Southeast Asia. MAM, March–April–May.

$^e$North Asia here extends poleward of 40°N and eastward of 43°E.

$^f$Central Africa here spans 22°S to the equator.

$^g$Middle East here covers 43°E–70°E and 12°N–40°N.

$^h$Australia here includes the islands of Oceania south of the equator.
from GOME NO$_2$ columns. As discussed in section 2.2, the error in the retrieved NO$_2$ column includes an absolute error of $1 \times 10^{13}$ molecules cm$^{-2}$, largely from the spectral fit, and a relative error of 42% from the AMF calculation. These errors were derived for individual retrievals, but we apply them to monthly mean fields to be conservative. If the retrieval errors are truly random, monthly mean errors would be 20% of the errors used here. Previous comparison of GEOS-CHEM model results with observations of the NO$_x$/NO$_2$ concentration ratio over the United States [Fiore et al., 2002] suggests that model error in the calculation of $\alpha$ adds another 30% error in quadrature to the mapping of NO$_2$ columns to NO$_x$ emissions. The resulting relative errors in top-down emissions, shown in Figure 5, are 1.5 over continental source regions and higher elsewhere. Remote regions with negative monthly mean NO$_2$ columns (reflecting errors in spectral fitting or subtraction of the stratospheric column) are assigned a very large error and hence contain no top-down information. We see from Figure 5 that top-down errors are comparable with bottom-up errors over industrial regions, and are much smaller than bottom-up errors over the southwestern United States, the Middle East, and Africa. Top-down errors exceed bottom-up errors over regions where retrieved columns are low and bottom-up emission estimates generally are low.

Table 2 provides regional top-down emissions. The values for North America are within 20% of the GOME top-down inventory calculated by Leue et al. [2001], but there are large differences for the rest of the world. Top-down emissions for Africa, Australia, and South America are about half those reported by Leue et al. [2001], and emissions for Europe and East Asia are 50–75% higher. Treatment of the AMF and the NO$_x$ lifetime likely explains the difference the two studies.

### 3.4. A Posteriori NO$_2$ Emissions

Figure 4 and Table 2 give the spatial distributions and regional totals of NO$_x$ emissions for the a posteriori inventory (computed from equation (1)). Figure 5 and Table 2 show that the a priori and top-down information contribute nearly equally to the a posteriori in the global mean. Table 2 also includes regional totals from the EDGAR emission inventory. Global annual surface NO$_x$ emissions in the a posteriori inventory are 37.7 Tg N yr$^{-1}$, not significantly different from the GEOS-CHEM a priori (36.4) or EDGAR 3.0 (36.6) values. Global annual errors are reduced from a factor of 2.0 for the a priori to 1.6 for the a posteriori. The monthly global a posteriori inventory is highly correlated with the a priori ($r^2 = 0.86, n = 35,568$), but there are some significant regional differences discussed below. We see from Figure 4 that there is little seasonal variation in a posteriori emissions from industrial regions of North America or Europe, where anthropogenic emissions (known to be largely aseasonal [National Acid Precipitation Assessment Program, 1991]) dominate. These regions show large seasonal variation in GOME NO$_2$ columns $\Omega_s$, but $\Omega_s/\tau_{NO_x}$ has in fact little seasonal variation, providing confidence in our approach.

A posteriori NO$_2$ emissions from the United States are 10% higher than the a priori inventory. A recent NO$_x$ emission estimate [Environmental Protection Agency, 2000] suggests that the a priori US emissions used here for 1996–1997 are biased low by 20%, consistent with the information from GOME data. Differences between the a posteriori and a priori inventories in the western United States are largest in summer, which could reflect an underestimate of soil NO$_x$ emissions in the a priori. Both a posteriori and EDGAR NO$_x$ emissions are higher than the a priori inventory for Europe, by up to 50% in the polluted Po Valley of Italy. A posteriori emissions from southern Europe are largest during summer, again suggesting a contribution from soil NO$_x$ emissions.

The largest regional percentage increases in NO$_x$ emissions from the a priori to the a posteriori are for the Middle East, northern Africa, South Africa, and Japan (Table 2). Figure 4 shows that much of the increase in NO$_x$ emissions for the Middle East is from Riyadh and Tehran which are underestimated by 50–100% throughout the year. The increase in NO$_x$ emissions for northern Africa represents a large percentage increase (25%) but a small absolute difference (0.8 Tg N yr$^{-1}$) over a large area. Most of the increase in NO$_x$ emissions is from the Sahel throughout the year. The algorithm of Yienger and Levy [1995] for soil NO$_x$ emissions used here as a priori may underestimate emissions from the Sahel by more than a factor of 2 [Le Roux et al., 1995; Davidson and Kingerlee, 1997]. South African emissions are underestimated by 25%, largely due to emissions from the industrial Highveld plateau. Emissions from Japan are underestimated by 35%.

Regional decreases for South Asia, South America, and central Africa between the a priori and the a posteriori emission inventories largely reflect biomass burning. Figure 4 shows local decreases of up to 50% for South Asia during March–May, and for South America and central Africa during July–October. In contrast, a posteriori emissions are higher over the northern African biomass burning region. Errors in the aerosol correction are insufficient to explain
these discrepancies [Martin, 2002]. Errors in NO₂ emission inventories for biomass burning arise from estimates of biomass burned and of the NOₓ emission factors (NOₓ emitted per kg of dry matter burned). The NOₓ emission factor for savannas is particularly uncertain (3.9 ± 2.4 g/kg) [Andreae and Merlet, 2001]. The NOₓ emission factors used in the a priori inventory are 1.1 g/kg for northern African savannas and 3.9 g/kg for other savannas [Staudt et al., 2003]. The a posteriori inventory suggests that NOₓ emission factors may be more uniform across different savanna types and closer to the low end recommended by Andreae and Merlet. Analyses of CO observations from the MOPITT instrument (C. L. Heald et al., Transpacific satellite and aircraft observations of Asian pollution, submitted to Journal of Geophysical Research, 2003) and from the TRACE-P aircraft mission [Palmer et al., 2003] are consistent with an overestimate of biomass burned in South Asia.

[29] A posteriori NOₓ emissions from South Asia are 20% lower than the a priori during the biomass burning season of March-May, and 10% lower during the rest of the year when anthropogenic emissions dominate. The GOME NO₂ columns would suggest a greater reduction of emissions during the rest of the year, but have limited weight relative to the a priori due to high relative errors (Figure 5). An anthropogenic NOₓ emission inventory for India [Garg et al., 2001] gives a national estimate of half the a priori inventory used here, but this inventory does not completely account for emissions from biomass burning, biofuels, and soils.

[30] As shown in Table 2, a posteriori regional emission inventories tend to fall between the EDGAR 3.0 and a priori inventories. Regional totals in the a posteriori inventory are higher than the EDGAR inventory for central Africa (55%), northern Asia (30%), and the United States (20%). A posteriori NOₓ emissions are lower than EDGAR emissions for Japan (40%), Canada (30%), and Europe (25%). The annual mean correlation r between EDGAR and the a posteriori inventory is 0.64 (n = 2964).

[31] Figure 5 shows that a posteriori errors are substantially lower than a priori errors throughout the world. Top-down information from GOME provides especially large error reductions over remote regions including Africa, the Middle East, South Asia, and the western United States. A posteriori errors over the eastern United States, Europe, and eastern China are typically less than 40%.

4. Conclusions

[32] We have retrieved global distributions of tropospheric NO₂ columns from solar backscatter measurements by the Global Ozone Monitoring Experiment (GOME) for September 1996–August 1997. Our retrieval improves upon that of Martin et al. [2002b] to account for aerosol effects. The air mass factor (AMF) calculation includes vertically resolved aerosol fields from a global aerosol model GOCART. Absorbing aerosols decrease the sensitivity of GOME to the NO₂ column over biomass burning regions by up to 40%. Scattering aerosols increase the sensitivity of GOME to NO₂ over the eastern United States by 5–10%.

[33] The AMF calculation in our retrieval uses relative vertical profiles (shape factors) of NO₂ from a global 3-D model of tropospheric chemistry (GEOS-CHEM). The GOME NO₂ retrievals carried out by other groups have assumed a globally uniform shape factor. We show that assuming a globally uniform shape factor introduces regional biases of up to 40% over industrial regions and a factor of 2 over remote regions.

[34] Retrieved NO₂ columns are lowest during summer, due to rapid loss by reaction with OH. Seasonal enhancements from biomass burning occur over northern Africa during DJF, over India during MAM, and over central Africa and South America during June-October. No enhancements are observed along ocean ship tracks, but this is not inconsistent with current ship emission inventories. No enhancements from lightning are observed over tropical continental regions, but again this is not inconsistent with current lightning emission estimates (most of the lightning NOₓ in the upper troposphere is present as NO at 1030 LT of the GOME measurement).

[35] We used the GOME retrievals of NO₂ columns to derive monthly top-down constraints for NOₓ emissions through inversion with the GEOS-CHEM model, and we combined these constraints with bottom-up inventories to obtain an optimized a posteriori estimate of NOₓ emissions from fossil fuels, biofuels, biomass burning, and soils. The optimization was weighted by the respective local errors in bottom-up and top-down estimates. We estimated the local errors in bottom-up industrial emissions by comparing the GEOS-CHEM and EDGAR 3.0 NOₓ emission inventories; these errors are on the order of 50%. Bottom-up errors over regions dominated by soils and biomass burning are taken to be a factor of 3. Top-down errors derived from formal error propagation are about 50% over continental source regions, comparable with the bottom-up errors. Our top-down estimates of NOₓ emissions over industrial source regions show little seasonal variation, as higher NO₂ columns in winter are offset by longer NOₓ lifetimes; this result is consistent with bottom-up knowledge and provides an important test of the top-down method.

[36] Global annual surface NOₓ emission in our a posteriori inventory is 37.7 Tg N yr⁻¹, which is not significantly different from the GEOS-CHEM a priori (36.4) or EDGAR 3.0 (36.6) values. About half of the information in the a posteriori inventory is from the top-down approach. Global annual errors are reduced from a factor of 2.0 for the a priori to 1.6 for the a posteriori. Monthly and spatial variation in a priori and a posteriori emissions are highly correlated ($r² = 0.86$, $n = 35,568$), but there are some significant regional differences. A posteriori NOₓ emissions are higher by 50–100% in the Po Valley, Tehran, and Riyadh urban areas, by 25–35% in Japan and South Africa, and by 10% in the United States. A priori soil NOₓ emissions appear to be too low over the Sahel, the western United States, and southern Europe. A priori soil NOₓ emissions are too high by up to 50% in biomass burning regions of India, central Africa, and Brazil. Noteworthy differences between the a posteriori and EDGAR inventories are that the EDGAR inventory is too low over central Africa (55%) and the United States (20%), but too high over Japan (40%) and Europe (25%). A posteriori errors on emission estimates are substantially lower than a priori errors throughout the world, especially over Africa, the Middle East, South Asia, and the western United States.

[37] Two major limitations of the current work are the coarse horizontal resolution of the GOME instrument and...
the lack of direct validation of the tropospheric NO\textsubscript{2} column product. Top-down emission inventories will be improved with the next generation of instruments featuring higher spatial resolution that enables more accurate cloud and aerosol corrections; the Scanning Imaging Absorption spectrometer for Atmospheric Chartography (SCIAMACHY) \cite{Bovensmann et al., 1999} on board ENVISAT and the Ozone Monitoring Instrument (OMI) on board Aura. Application of daily retrieved aerosol fields would increase the accuracy of the aerosol correction. Inversion of NO\textsubscript{2} emissions at a higher resolution will require a more formal inverse analysis to account for transport. The higher resolution measurements may enable observation of ship emissions over oceans. Coincident limb measurements from both platforms will enable better separation of the stratosphere and troposphere. Direct validation of satellite observations with in situ measurements and comparison with high-resolution, customized bottom-up inventories will be needed for further progress.

Appendix A: GEOS-CHEM Model Description

[38] The determination of NO\textsubscript{2} emissions from GOME observations of tropospheric NO\textsubscript{2} columns requires independent information on the relationship of NO\textsubscript{2} columns to NO\textsubscript{2} emissions. A global 3-D model of tropospheric chemistry is the best source of this information. The GEOS-CHEM model was initially described by \textit{Bey et al.} \cite{2001a}. Subsequent improvements are described by \textit{Martin et al.} \cite{2002b, 2003}. We use here GEOS-CHEM version 4.26 (http://www-as.harvard.edu/chemistry/trop/geo/) with improved treatment of aerosol photochemical effects and a few other updates as described by \textit{Martin et al.} \cite{2003}. The model is driven by assimilated meteorological data from the Goddard Earth Observing System (GEOS) of the NASA Data Assimilation Office. The meteorological data include 3-D fields updated every 3 hours for surface fluxes and mixing depths, and every 6 hours for other variables. The NO\textsubscript{2} emissions inventory is discussed in section 3.2. We use for this study the GEOS data for 1996–1997, available with the original horizontal resolution. For this study the GEOS data for 1996–1997, available with the original horizontal resolution.

[39] The GEOS-CHEM model includes a detailed simulation of tropospheric O\textsubscript{3},NO\textsubscript{2},hydrocarbon chemistry, originally described by \textit{Horowitz et al.} \cite{1998} and updated as described by several people \textit{Bey et al.}, 2001a; \textit{Fiore et al.}, 2002; \textit{Martin et al.}, 2002a, 2003. Reactions in aerosols, including N\textsubscript{2}O\textsubscript{5} hydrolysis (reaction probability 0.1), are described by \textit{Jacob} \cite{2000}. The chemical evolution of about 120 species is computed with a Gear solver \cite{Jacobson and Turco, 1994}. Photolysis frequencies are computed using the Fast-J radiative transfer algorithm \cite{Wild et al., 2000}, which includes Rayleigh scattering as well as Mie scattering by clouds and aerosols. Aerosol fields affecting radiation and heterogeneous chemistry are 3-D monthly means from a 1996–1997 simulation with the GOCART model \cite{Chin et al., 2000a, 2002; Ginoux et al., 2001}, which uses the same GEOS meteorological fields and transport algorithms as GEOS-CHEM.

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